

Package **AquaEnv**: an Aquatic modelling Environment in R

Andreas F. Hofmann

Centre for Estuarine and Marine Ecology

Netherlands Institute of Ecology

The Netherlands

Abstract

AquaEnv is an integrated development toolbox for aquatic chemical model generation focused on (ocean) acidification and CO₂ air-water exchange.

- It contains all elements necessary to model the pH, the related CO₂ air-water exchange, as well as aquatic acid-base chemistry in general for an arbitrary marine, estuarine or freshwater system. Also chemical batches can be modelled.
- Next to the routines necessary to calculate desired information, **AquaEnv** also contains a suite of tools to visualize this information.
- Furthermore, **AquaEnv** can not only be used to build dynamic models of aquatic systems, but it can also serve as a simple desktop tool for the experimental aquatic chemist to generate and visualize all possible derived information from a set of measurements with one single easy to use R function.
- Additionally, the sensitivity of the system to variations in the input variables can be visualized.
- **AquaEnv** also contains a number of example “applications” that make use of the aquatic modelling toolbox that **AquaEnv** provides:
 - a theoretical titration simulator
 - and a routine to determine total alkalinity ([TA]), the total dissolved inorganic carbon concentration ($[\sum \text{CO}_2]$), as well as additionally the electrode standard potential (E_0) and the first dissociation constant of the carbonate system ($K_{\text{CO}_2}^*$) from titration data.

Keywords: aquatic modelling, pH, pH scales, dissolved inorganic carbon, total alkalinity, total alkalinity curve fitting, theoretical titration, revelle factor, omega, solubility products, CO₂, ocean acidification, estuaries, carbonate system, seawater, R.

Contents

1	Introduction	4
2	The elements of an object of class <i>aquaenv</i>	6
3	Using AquaEnv	10
3.1	Basic features	10
3.1.1	calling the “K” functions directly	10
3.1.2	Minimal <i>aquaenv</i> definition	11
3.1.3	Defining the complete <i>aquaenv</i> system in different ways	14
3.1.4	Calculating $[\sum \text{CO}_2]$	17
3.1.5	Cloning an object of class <i>aquaenv</i>	18
3.1.6	Preparing input variables	19
3.1.7	Vectors as input variables	19
3.1.8	Calculating $[\sum \text{CO}_2]$ from input vectors	22
3.1.9	Conversion from and to a dataframe	22
3.1.10	Converting elements in an object of class <i>aquaenv</i>	23
3.1.11	Quantities needed for explicit pH modelling	23
3.2	The <code>plot.aquaenv</code> function	24
3.3	Using objects of class <i>aquaenv</i> in dynamic models	26
3.3.1	Ordinary dynamic models	26
3.3.2	Models using the explicit pH modelling approach	28
3.3.2.1	In one single model	28
3.3.2.2	In three separate models	33
3.3.2.2.1	The implicit pH modelling approach	33
3.3.2.2.2	The explicit pH modelling approach	34
3.3.2.2.3	The fractional stoichiometric approach	35
3.4	Titration simulation: the function <code>titration</code>	36
3.4.1	Titration with HCl	37
3.4.2	Titration with NaOH	45
3.4.3	Titration with a titrant with high concentrations and a large sample volume - classical Bjerrum plots	46
3.5	Calculating information from titration curves: the function <code>TAfit</code>	48
3.5.1	A little theory	48
3.5.2	Determining $[\text{TA}]$ and $[\sum \text{CO}_2]$ by non linear curve fitting	52
3.5.2.1	Proof of concept	52
3.5.2.2	Test with generated data from literature	57
3.5.2.2.1	Does the salinity correction (<code>S_titrant</code>) matter?	58
3.5.2.2.2	Does fitting <code>K_CO2</code> as well improve the fit?	61

4	Extending AquaEnv	65
A	Abbreviations for references used throughout the code and in the helpfiles	69
B	References for the elements of an object of class <i>aquaenv</i>	70

1 Introduction

AquaEnv is a toolbox for aquatic modelling that serves several purposes

- It provides functions to calculate the stoichiometric equilibrium constants (K^*) for key acid base systems in natural seawater, the Henry's constants (K_0), as well as the solubility products (K_{sp}) for calcite and aragonite. This functionality is provided via the functions `K_CO2`, `K_HCO3`, `K_BOH3`, `K_W`, `K_HSO4`, `K_HF`, `K_NH4`, `K_H2S`, `K_H3PO4`, `K_H2PO4`, `K_HPO4`, `K_SiOH4`, `K_SiOOH3`, `KO_CO2`, `KO_O2`, `Ksp_aragonite`, and `Ksp_calcite`.
- It is designed to make its use as easy as possible: all the information that can be calculated from the set of parameters known of a system or sample can be obtained by one single function: `aquaenv`. This function returns a list of class *aquaenv* that contains next to the input parameters
 - the chlorinity, the ionic strength, $[\sum B(OH)_3]$, $[\sum H_2SO_4]$, $[\sum HF]$, $[Cl^-]$, $[Cl^-]$, $[\sum Br]$, $[Na^+]$, $[Mg^{2+}]$, $[Ca^{2+}]$, $[K^+]$, $[Sr^{2+}]$ calculated from salinity as given in [DOE \(1994\)](#) (Please note that if values for $[\sum B(OH)_3]$, $[\sum H_2SO_4]$, $[\sum HF]$ are given as input parameters, these parameters are used and not the ones calculated from salinity.)
 - the gauge pressure p (total pressure minus atmospheric pressure [Feistel 2008](#)) either given as input variable, or calculated from depth (according to [Fofonoff and Millard 1983](#)), or calculated from the total pressure P and the atmospheric pressure P_a , both of which can be given as input variables and are also stored in an object of class *aquaenv*
 - the seawater density calculated from temperature and salinity as given by [Millero and Poisson \(1981\)](#)
 - a set of conversion factors to convert between different pH scales ([Dickson 1984](#); [Zeebe and Wolf-Gladrow 2001](#)) and between mol/kg-H₂O and mol/kg-solution (inferred from [Roy, Roy, Vogel, PorterMoore, Pearson, Good, Millero, and Campbell \(1993b\)](#) and [DOE \(1994\)](#))
 - the Henry's constants for CO₂ ([Weiss 1974](#)) and for O₂ (inferred from [Weiss 1970](#)) calculated from temperature and salinity as well as the associated saturation concentrations of CO₂ and O₂.
 - the ion product of water ([Millero 1995](#)), the stoichiometric equilibrium constants of HSO₄⁻ ([Dickson 1990a](#)), HF ([Dickson and Riley 1979a](#)), CO₂ ([Roy et al. 1993b](#)), HCO₃⁻ ([Roy et al. 1993b](#)), B(OH)₃ ([Dickson 1990a](#)), NH₄⁺ ([Millero, Yao, and Aicher 1995](#)), H₂S ([Millero 1995](#)), H₃PO₄ ([Millero 1995](#)), H₂PO₄⁻ ([Millero 1995](#)), HPO₄²⁻ ([Millero 1995](#)), SiOH₄ ([Millero, Plese, and Fernandez 1988](#)), SiOOH₃⁻ ([Wischmeyer, Del Amo, Brzezinski, and Wolf-Gladrow 2003](#)), HNO₂ ([Riordan, Minogue, Healy, O'Driscoll, and Sodeau 2005](#)), HNO₃, H₂SO₄ ([Atkins 1996](#)), HS ([Atkins 1996](#)) mostly calculated as functions of temperature and salinity and pressure corrected according to [Millero \(1995\)](#).
 - the solubility products of calcite and aragonite ([Mucci 1983](#)) as well as the associated Ω 's if a full speciation is calculated (see below)
 - the fugacity of CO₂ - if a full speciation is calculated (see below)

- if $[\sum \text{CO}_2]$ and pH are given $[\text{TA}]$ is calculated, if $[\sum \text{CO}_2]$ and $[\text{TA}]$ are given pH is calculated, if $[\sum \text{CO}_2]$ and $[\text{CO}_2]$ or fCO_2 are given, pH and $[\text{TA}]$ are calculated.
 - if either one of the pairs pH and $[\text{CO}_2]$ or fCO_2 , pH and $[\text{TA}]$, or $[\text{TA}]$ and $[\text{CO}_2]$ or fCO_2 is given, $[\sum \text{CO}_2]$ is calculated
 - if sufficient information is given and the flag `speciation=TRUE` is set, a full speciation of $[\sum \text{CO}_2]$, $[\sum \text{NH}_4]$, $[\sum \text{H}_2\text{S}]$, $[\sum \text{HNO}_3]$, $[\sum \text{HNO}_2]$, $[\sum \text{H}_3\text{PO}_4]$, $[\sum \text{Si}(\text{OH})_4]$, $[\sum \text{B}(\text{OH})_3]$, $[\sum \text{H}_2\text{SO}_4]$, $[\sum \text{HF}]$, as well as water itself is calculated
 - if the flag `revelle=TRUE` is set, the *revelle factor* (Zeebe and Wolf-Gladrow 2001) is calculated. item if the flag `revelle=TRUE` is set, all necessary quantities for the explicit “direct substitution approach” (DSA) to pH modelling as given in Hofmann, Meysman, Soetaert, and Middelburg (2008a) are calculated. These are the buffer factor (the partial derivative of $[\text{TA}]$ with respect to $[\text{H}^+]$) and the partial derivatives of $[\text{TA}]$ with respect to the other total quantities. Furthermore, the partial derivatives of $[\text{TA}]$ with respect to changes in the equilibrium constants (K^*), multiplied with the partial derivatives of the equilibrium constants with respect to their variables needed for the DSA with time variable equilibrium constants as described in Hofmann, Meysman, Soetaert, and Middelburg (2009) are calculated. Finally, the ionization fractions as defined by Stumm and Morgan (1996) and used in Hofmann, Middelburg, Soetaert, Wolf-Gladrow, and Meysman (2008b) are calculated for the full speciation.
- Input for `aquaenv` has to be supplied in standard SI units, the free proton pH scale and in molinity¹ (mol/kg-solution). Conversion of input parameters to this necessary units and pH scale can be done with the generic function `convert`.
 - The information created with `aquaenv` is also supplied in standard SI units and in molinity. All elements of an object of class `aquaenv` of a certain unit or pH scale can be converted into other units or pH scales with the function `convert` as well.
 - One can use input vectors of salinity *S*, temperature *t*, or gauge pressure *p* (as well as total pressure *P* and depth *d*) for `aquaenv` to obtain vectors of all calculated information as function of the input vector. This can be visualized in a large variety of ways using the `plot` function specially defined for objects pf class `aquaenv`.
 - Objects of class `aquaenv` can be used in dynamic models to define the state of the system in each timestep of the numerical integration (done e.g. with `deSolve`). with the function `aquaenv` and the flag `from.data.frame=TRUE` it is possible to convert output of those dynamic models into objects pf class `aquaenv` which allows the user to use the whole suite of visualisation tools that is provided by the function `plot` in **AquaEnv**.
 - As mentioned above Hofmann *et al.* (2008a), Hofmann *et al.* (2009), and Hofmann *et al.* (2008b) describe methods for an “explicit” pH modelling which allows for the quantification of the influences of kinetically modelled processes on the pH. Objects pf class `aquaenv` provide all needed quantities (partial derivatives of $[\text{TA}]$, ionization fractions, etc.) to employ both of those methods in dynamic models. Furthermore,

¹Note that it is not sufficient to give a gravimetric concentration in mol/kg since there is a substantial difference between mol/kg-H₂O (molality) and mol/kg-solution (molinity).

AquaEnv provides the functionality to cumulatively plot the obtained influences on the pH.

- As an example of how to use the aquatic chemical toolbox that is provided by **AquaEnv**, two applications are provided
 - The function `titration`: creates theoretical titrations which can be used e.g. to create bjerrum plots with the function `plot.aquaenv` in **AquaEnv**.
 - The function `TAfit`: a routine based on a method in DOE (1994) that makes use of that theoretical titration function and allows for determining total alkalinity ([TA]), the total dissolved inorganic carbon concentration ($[\sum \text{CO}_2]$), as well as additionally the electrode standard potential (E_0) and the first dissociation constant of the carbonate system ($K_{\text{CO}_2}^*$) using the Levenberg-Marquart algorithm (least squares optimization procedure) as provided in the R package **minpack.lm**.

2 The elements of an object of class *aquaenv*

The function `aquaenv`, the central function of **AquaEnv**, returns an object of class *aquaenv*. This object is a list of different elements which can be accessed with the `$` character or with the `[[` operator

```
> test <- aquaenv(35, 10)
> test$t
```

```
[1] 10
attr(,"unit")
[1] "deg C"
```

```
> test[["t"]]
```

```
[1] 10
attr(,"unit")
[1] "deg C"
```

Maximally, i.e., if the enough input data is supplied to define the pH of the system and the flags `speciation`, `dsa`, and `revelle` are TRUE while the flag `skeleton` is FALSE, an object of class *aquaenv* contains the following elements

element	unit	explanation
S	“psu” (no unit)	salinity
t	°C	temperature
p	bar	gauge pressure (total pressure minus atmospheric pressure Feistel 2008)
T	K	absolute temperature
Cl	‰	chlorinity
I	mol/kg-H ₂ O	ionic strength

P	bar	total pressure
Pa	bar	atmospheric pressure
d	m	depth
density	kg/m ³	(seawater) density
SumCO2	mol/kg-soln	[$\sum \text{CO}_2$], total dissolved inorganic carbon concentration
SumNH4	mol/kg-soln	[$\sum \text{NH}_4^+$], total ammonium concentration
SumH2S	mol/kg-soln	[$\sum \text{H}_2\text{S}$], total sulfide concentration
SumHNO3	mol/kg-soln	[$\sum \text{HNO}_3$], total nitrate concentration
SumHNO2	mol/kg-soln	[$\sum \text{HNO}_2$], total nitrite concentration
SumH3PO4	mol/kg-soln	[$\sum \text{H}_3\text{PO}_4$], total phosphate concentration
SumSiOH4	mol/kg-soln	[$\sum \text{Si}(\text{OH})_4$], total silicate concentration
SumBOH3	mol/kg-soln	[$\sum \text{B}(\text{OH})_3$], total borates concentration
SumH2SO4	mol/kg-soln	[$\sum \text{H}_2\text{SO}_4$], total sulfate concentration
SumHF	mol/kg-soln	[$\sum \text{HF}$], total fluoride concentration
Br	mol/kg-soln	[Br ⁻], bromide concentration
ClConc	mol/kg-soln	[Cl ⁻], chloride concentration
Na	mol/kg-soln	[Na ⁺], sodium concentration
Mg	mol/kg-soln	[Mg ²⁺], magnesium concentration
Ca	mol/kg-soln	[Ca ²⁺], calcium concentration
K	mol/kg-soln	[K ⁺], potassium concentration
Sr	mol/kg-soln	[Sr ²⁺], strontium concentration
molal2molin	(mol/kg-soln)/(mol/kg-H2O)	concentration conversion factor: from molality to molinity
free2tot	-	pH conversion factor: free scale to total scale
free2sws	-	pH conversion factor: free scale to seawater scale
tot2free	-	pH conversion factor: total scale to free scale
tot2sws	-	pH conversion factor: total scale to seawater scale
sws2free	-	pH conversion factor: seawater scale to free scale
sws2tot	-	pH conversion factor: seawater scale to total scale
K0_CO2	mol/(kg-soln*atm)	Henry's constant for CO ₂
K0_O2	mol/(kg-soln*atm)	Henry's constant for O ₂
fCO2atm	atm	atmospheric fugacity of CO ₂
fO2atm	atm	atmospheric fugacity of O ₂
CO2_sat	mol/kg-soln	CO ₂ saturation concentration at an atmospheric fugacity of fCO2atm
O2_sat	mol/kg-soln	O ₂ saturation concentration at an atmospheric fugacity of fO2atm
K_W	(mol/kg-soln) ² , free pH scale	stoichiometric equilibrium ion product of H ₂ O: $K_W^* = [\text{H}^+][\text{OH}^-]$
K_HSO4	mol/kg-soln, free pH scale	stoichiometric equilibrium constant $K_{\text{HSO}_4}^* = [\text{H}^+][\text{SO}_4^{2-}]/[\text{HSO}_4^-]$
K_HF	mol/kg-soln, free pH scale	stoichiometric equilibrium constant $K_{\text{HF}}^* = [\text{H}^+][\text{F}^-]/[\text{HF}]$
K_CO2	mol/kg-soln, free pH scale	stoichiometric equilibrium constant $K_{\text{CO}_2}^* = [\text{H}^+][\text{HCO}_3^-]/[\text{CO}_2]$

K_HCO3	mol/kg-soln, free pH scale	stoichiometric equilibrium constant $K_{\text{HCO}_3^-}^* = [\text{H}^+][\text{CO}_3^{2-}]/[\text{HCO}_3^-]$
K_BOH3	mol/kg-soln, free pH scale	stoichiometric equilibrium constant $K_{\text{B(OH)}_3}^* = [\text{H}^+][\text{B(OH)}_4^-]/[\text{B(OH)}_3]$
K_NH4	mol/kg-soln, free pH scale	stoichiometric equilibrium constant $K_{\text{NH}_4^+}^* = [\text{H}^+][\text{NH}_3]/[\text{NH}_4^+]$
K_H2S	mol/kg-soln, free pH scale	stoichiometric equilibrium constant $K_{\text{H}_2\text{S}}^* = [\text{H}^+][\text{HS}^-]/[\text{H}_2\text{S}]$
K_H3PO4	mol/kg-soln, free pH scale	stoichiometric equilibrium constant $K_{\text{H}_3\text{PO}_4}^* = [\text{H}^+][\text{H}_2\text{PO}_4^-]/[\text{H}_3\text{PO}_4]$
K_H2PO4	mol/kg-soln, free pH scale	stoichiometric equilibrium constant $K_{\text{H}_2\text{PO}_4^-}^* = [\text{H}^+][\text{HPO}_4^{2-}]/[\text{H}_2\text{PO}_4^-]$
K_HP04	mol/kg-soln, free pH scale	stoichiometric equilibrium constant $K_{\text{HPO}_4^{2-}}^* = [\text{H}^+][\text{PO}_4^{3-}]/[\text{HPO}_4^{2-}]$
K_SiOH4	mol/kg-soln, free pH scale	stoichiometric equilibrium constant $K_{\text{Si(OH)}_4}^* = [\text{H}^+][\text{SiO(OH)}_3^-]/[\text{Si(OH)}_4]$
K_Si00H3	mol/kg-soln, free pH scale	stoichiometric equilibrium constant $K_{\text{SiO(OH)}_3^-}^* = [\text{H}^+][\text{SiO}_2(\text{OH})_2^{2-}]/[\text{SiO(OH)}_3^-]$
K_HNO2	mol/kg-soln; mol/kg-H2O; mol/l	approximate value for equilibrium constant $K_{\text{HNO}_2}^* = [\text{H}^+][\text{NO}_2^-]/[\text{HNO}_2]$
K_HNO3	mol/kg-soln; mol/kg-H2O; mol/l	approximate value for equilibrium constant $K_{\text{HNO}_3}^* = [\text{H}^+][\text{NO}_3^-]/[\text{HNO}_3]$
K_H2S04	mol/kg-soln; mol/kg-H2O; mol/l	approximate value for equilibrium constant $K_{\text{H}_2\text{SO}_4}^* = [\text{H}^+][\text{HSO}_4^-]/[\text{H}_2\text{SO}_4]$
K_HS	mol/kg-soln; mol/kg-H2O; mol/l	approximate value for equilibrium constant $K_{\text{HS}^-}^* = [\text{H}^+][\text{S}^{2-}]/[\text{HS}^-]$
Ksp_calcite	(mol/kg-soln) ²	stoichiometric equilibrium solubility product of calcite $K_{\text{sp}}^*_{\text{cal}} = [\text{Ca}^{2+}][\text{CO}_3^{2-}]$
Ksp_aragonite	(mol/kg-soln) ²	stoichiometric equilibrium solubility product of aragonite $K_{\text{sp}}^*_{\text{ara}} = [\text{Ca}^{2+}][\text{CO}_3^{2-}]$
TA	mol/kg-soln	[TA], total alkalinity
pH	-, free scale	pH
fCO2	atm,	fugacity of CO ₂ in the water (i.e. in a small volume of air equilibrated with the water)
C02	mol/kg-soln	[CO ₂]
HC03	mol/kg-soln	[HCO ₃ ⁻]
C03	mol/kg-soln	[CO ₃ ²⁻]
B0H3	mol/kg-soln	[B(OH) ₃]
B0H4	mol/kg-soln	[B(OH) ₄ ⁻]
OH	mol/kg-soln	[OH ⁻]
H3P04	mol/kg-soln	[H ₃ PO ₄]
H2P04	mol/kg-soln	[H ₂ PO ₄ ⁻]
HP04	mol/kg-soln	[HPO ₄ ²⁻]

P04	mol/kg-soln	$[\text{PO}_4^{3-}]$
SiOH4	mol/kg-soln	$[\text{Si}(\text{OH})_4]$
Si00H3	mol/kg-soln	$[\text{SiO}(\text{OH})_3^-]$
Si02OH2	mol/kg-soln	$[\text{SiO}_2(\text{OH})_2^{2-}]$
H2S	mol/kg-soln	$[\text{H}_2\text{S}]$
HS	mol/kg-soln	$[\text{HS}^-]$
S2min	mol/kg-soln	$[\text{S}^{2-}]$
NH4	mol/kg-soln	$[\text{NH}_4^+]$
NH3	mol/kg-soln	$[\text{NH}_3]$
H2SO4	mol/kg-soln	$[\text{H}_2\text{SO}_4]$
HSO4	mol/kg-soln	$[\text{HSO}_4^-]$
SO4	mol/kg-soln	$[\text{SO}_4^{2-}]$
HF	mol/kg-soln	$[\text{HF}]$
F	mol/kg-soln	$[\text{F}^-]$
HN03	mol/kg-soln	$[\text{HNO}_3]$
N03	mol/kg-soln	$[\text{NO}_3^-]$
HN02	mol/kg-soln	$[\text{HNO}_2]$
N02	mol/kg-soln	$[\text{NO}_2^-]$
omega_calcite	-	saturation state Ω with respect to calcite
omega_aragonite	-	saturation state Ω with respect to aragonite
revelle	-	Revelle factor
c1	-	ionization fraction $c_1 = [\text{CO}_2]/[\sum \text{CO}_2]$
c2	-	ionization fraction $c_2 = [\text{HCO}_3^-]/[\sum \text{CO}_2]$
c3	-	ionization fraction $c_3 = [\text{CO}_3^{2-}]/[\sum \text{CO}_2]$
dTAdSumCO2	-	$\frac{\partial[\text{TA}]}{\partial \sum \text{CO}_2}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots)$
b1	-	ionization fraction $b_1 = [\text{B}(\text{OH})_3]/[\sum \text{B}(\text{OH})_3]$
b2	-	ionization fraction $b_2 = [\text{B}(\text{OH})_4^-]/[\sum \text{B}(\text{OH})_3]$
dTAdSumBOH3	-	$\frac{\partial[\text{TA}]}{\partial \sum \text{B}(\text{OH})_3}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots)$
so1	-	ionization fraction $so_1 = [\text{H}_2\text{SO}_4]/[\sum \text{H}_2\text{SO}_4]$
so2	-	ionization fraction $so_2 = [\text{HSO}_4^-]/[\sum \text{H}_2\text{SO}_4]$
so3	-	ionization fraction $so_3 = [\text{SO}_4^{2-}]/[\sum \text{H}_2\text{SO}_4]$
dTAdSumH2SO4	-	$\frac{\partial[\text{TA}]}{\partial \sum \text{H}_2\text{SO}_4}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots)$
f1	-	ionization fraction $f_1 = [\text{HF}]/[\sum \text{HF}]$
f2	-	ionization fraction $f_1 = [\text{F}^-]/[\sum \text{HF}]$
dTAdSumHF	-	$\frac{\partial[\text{TA}]}{\partial \sum \text{HF}}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots)$
p1	-	ionization fraction $p_1[\text{H}_3\text{PO}_4]/[\sum \text{H}_3\text{PO}_4]$
p2	-	ionization fraction $p_2[\text{H}_2\text{PO}_4^-]/[\sum \text{H}_3\text{PO}_4]$
p3	-	ionization fraction $p_3[\text{HPO}_4^{2-}]/[\sum \text{H}_3\text{PO}_4]$
p4	-	ionization fraction $p_4[\text{PO}_4^{3-}]/[\sum \text{H}_3\text{PO}_4]$
dTAdSumH3PO4	-	$\frac{\partial[\text{TA}]}{\partial \sum \text{H}_3\text{PO}_4}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots)$

si1	-	ionization fraction $si_1 = [\text{Si}(\text{OH})_4] / [\sum \text{Si}(\text{OH})_4]$
si2	-	ionization fraction $si_2 = [\text{SiO}(\text{OH})_3] / [\sum \text{Si}(\text{OH})_4]$
si3	-	ionization fraction $si_3 = [\text{SiO}_2(\text{OH})_2] / [\sum \text{Si}(\text{OH})_4]$
dTAdSumSumSiOH4	-	$\frac{\partial[\text{TA}]}{\partial \sum \text{Si}(\text{OH})_4}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots)$
s1	-	ionization fraction $s_1 = [\text{H}_2\text{S}] / [\sum \text{H}_2\text{S}]$
s2	-	ionization fraction $s_2 = [\text{HS}^-] / [\sum \text{H}_2\text{S}]$
s3	-	ionization fraction $s_3 = [\text{S}^{2-}] / [\sum \text{H}_2\text{S}]$ Note that we do assume S^{2-} does exist. However, s_3 is very small.
dTAdSumH2S	-	$\frac{\partial[\text{TA}]}{\partial \sum \text{H}_2\text{S}}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots)$
n1	-	ionization fraction $n_1 = [\text{NH}_4^+] / [\sum \text{NH}_4^+]$
n2	-	ionization fraction $n_2 = [\text{NH}_3] / [\sum \text{NH}_4^+]$
dTAdSumNH4	-	$\frac{\partial[\text{TA}]}{\partial \sum \text{NH}_4^+}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots)$
na1	-	ionization fraction $na_1 = [\text{HNO}_3] / [\sum \text{HNO}_3]$
na2	-	ionization fraction $na_2 = [\text{NO}_3^-] / [\sum \text{HNO}_3]$
dTAdSumHNO3	-	$\frac{\partial[\text{TA}]}{\partial \sum \text{HNO}_3}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots)$
ni1	-	ionization fraction $ni_1 = [\text{HNO}_2] / [\sum \text{HNO}_2]$
ni2	-	ionization fraction $ni_2 = [\text{NO}_2^-] / [\sum \text{HNO}_2]$
dTAdSumHNO2	-	$\frac{\partial[\text{TA}]}{\partial \sum \text{HNO}_2}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots)$
dTAdH	-	$\frac{\partial[\text{TA}]}{\partial [\text{H}^+]}$: buffer factor with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots)$
dTAdKdKdS	-	$\sum_i \frac{\partial[\text{TA}]}{\partial K_i^*} \frac{\partial K_i^*}{\partial S}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots, K_i^*)$
dTAdKdKdT	-	$\sum_i \frac{\partial[\text{TA}]}{\partial K_i^*} \frac{\partial K_i^*}{\partial T}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots, K_i^*)$
dTAdKdKdp	-	$\sum_i \frac{\partial[\text{TA}]}{\partial K_i^*} \frac{\partial K_i^*}{\partial p}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots, K_i^*)$
dTAdKdKdSumH2SO4	-	$\sum_i \frac{\partial[\text{TA}]}{\partial K_i^*} \frac{\partial K_i^*}{\partial [\sum \text{H}_2\text{SO}_4]}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots, K_i^*)$
dTAdKdKdSumHF	-	$\sum_i \frac{\partial[\text{TA}]}{\partial K_i^*} \frac{\partial K_i^*}{\partial [\sum \text{HF}]}$ with $[\text{TA}] = f([\text{H}^+], [\sum \text{CO}_2], \dots, K_i^*)$

For elements that are calculated according to certain literature references, those references are given in appendix B.

3 Using AquaEnv

3.1 Basic features

3.1.1 calling the “K” functions directly

The elements `K_CO2`, `K_HCO3`, `K_BOH3`, `K_W`, `K_HSO4`, `K_HF`, `K_NH4`, `K_H2S`, `K_H3PO4`, `K_H2PO4`, `K_HP04`, `K_SiOH4`, `K_SiOOH3`, `KO_CO2`, `KO_O2`, `Ksp_aragonite`, and `Ksp_calcite` can be calculated directly, without creating an object of class *aquaenv*. This is done via functions that bear the same name as those elements

```
> K_CO2(15, 30)

[1] 1.002156e-06
attr("unit")
[1] "mol/kg-soln"
attr("pH scale")
[1] "free"

> KO_CO2(15, 30)

[1] 0.02780196
attr("unit")
[1] "mol/(kg-soln*atm)"

> Ksp_calcite(15, 30, 100)

[1] 1.721545e-07
attr("unit")
[1] "(mol/kg-soln)^2"
```

3.1.2 Minimal *aquaenv* definition

Minimally, an object of class *aquaenv* can be defined with just a temperature and salinity value

```
> ae <- aquaenv(S = 30, t = 15)
> ae$K_CO2

[1] 9.089094e-07
attr("unit")
[1] "mol/kg-soln"
attr("pH scale")
[1] "free"
```

Optionally, a the pressure (here the gauge pressure) can be given. As in the above case, the returned object of class *aquaenv* then contains a standard set of elements as shown by the `names` command.

```

> ae <- aquaenv(S = 30, t = 15, p = 10)
> ae$Ksp_calcite

[1] 3.643728e-07
attr(,"unit")
[1] "(mol/kg-soln)^2"

> names(ae)

[1] "S"          "t"          "p"          "T"
[5] "Cl"         "I"          "p"          "Pa"
[9] "d"          "density"    "SumCO2"     "SumNH4"
[13] "SumH2S"     "SumHNO3"    "SumHNO2"    "SumH3PO4"
[17] "SumSiOH4"   "SumBOH3"    "SumH2SO4"    "SumHF"
[21] "Br"         "ClConc"     "Na"          "Mg"
[25] "Ca"         "K"          "Sr"          "molal2molin"
[29] "free2tot"   "free2sws"   "tot2free"    "tot2sws"
[33] "sws2free"   "sws2tot"    "K0_CO2"      "K0_O2"
[37] "fCO2atm"    "fO2atm"     "CO2_sat"     "O2_sat"
[41] "K_W"        "K_HSO4"     "K_HF"        "K_CO2"
[45] "K_HCO3"     "K_BOH3"     "K_NH4"       "K_H2S"
[49] "K_H3PO4"    "K_H2PO4"    "K_HPO4"      "K_SiOH4"
[53] "K_SiOOH3"   "K_HNO2"     "K_HNO3"      "K_H2SO4"
[57] "K_HS"       "Ksp_calcite" "Ksp_aragonite"

```

The pressure can also be given via the total pressure or the water depth. The atmospheric pressure can be given as well (e.g. for the case of a mountain lake). Furthermore, if the depth is given, the latitude can be specified (default is 0 degrees).

```

> ae <- aquaenv(S = 30, t = 15, p = 10)
> ae$p

[1] 10
attr(,"unit")
[1] "bar"

> ae$P

[1] 11.01325
attr(,"unit")
[1] "bar"

> ae <- aquaenv(S = 30, t = 15, P = 10)
> ae$p

[1] 8.98675
attr(,"unit")
[1] "bar"

```

```
> ae$P

[1] 10
attr(,"unit")
[1] "bar"

> ae <- aquaenv(S = 30, t = 15, P = 10, Pa = 0.5)
> ae$p

[1] 9.5
attr(,"unit")
[1] "bar"

> ae$P

[1] 10
attr(,"unit")
[1] "bar"

> ae <- aquaenv(S = 30, t = 15, d = 100)
> ae$p

[1] 10.05769
attr(,"unit")
[1] "bar"

> ae$P

[1] 11.07094
attr(,"unit")
[1] "bar"

> ae <- aquaenv(S = 30, t = 15, d = 100, lat = 51)
> ae$p

[1] 10.08985
attr(,"unit")
[1] "bar"

> ae$P

[1] 11.10310
attr(,"unit")
[1] "bar"
```

A minimal set of elements in an object of class *aquaenv* can be obtained by setting the flag `skeleton` to `TRUE`

```
> ae <- aquaenv(S = 30, t = 15, p = 10, skeleton = TRUE)
> names(ae)
```

```
[1] "S"      "t"      "p"      "T"      "Cl"     "I"
[7] "P"      "Pa"     "d"      "density" "SumCO2" "SumNH4"
[13] "SumH2S" "SumHNO3" "SumHNO2" "SumH3PO4" "SumSiOH4" "SumBOH3"
[19] "SumH2SO4" "SumHF" "K_W" "K_HSO4" "K_HF" "K_CO2"
[25] "K_HCO3" "K_BOH3" "K_NH4" "K_H2S" "K_H3PO4" "K_H2PO4"
[31] "K_HPO4" "K_SiOH4" "K_SiOOH3" "K_HNO2" "K_HNO3" "K_H2SO4"
[37] "K_HS"
```

3.1.3 Defining the complete *aquaenv* system in different ways

If enough information is given to define a complete speciation, i.e. either one of the pairs SumCO2 and pH, SumCO2 and TA, SumCO2 and CO2, or SumCO2 and fCO2, a full *aquaenv* system can be defined.

```
> S <- 30
> t <- 15
> p <- 10
> SumCO2 <- 0.002
> pH <- 8
> TA <- 0.002142233
> fCO2 <- 0.0005272996
> CO2 <- 2.031241e-05
> ae <- aquaenv(S, t, p, SumCO2 = SumCO2, pH = pH)
> ae$TA
```

```
[1] 0.002142243
attr(,"unit")
[1] "mol/kg-soln"
```

```
> ae <- aquaenv(S, t, p, SumCO2 = SumCO2, TA = TA)
> ae$pH
```

```
[1] 7.999976
attr(,"pH scale")
[1] "free"
```

```
> ae <- aquaenv(S, t, p, SumCO2 = SumCO2, CO2 = CO2)
> ae$pH
```

```
[1] 7.999975
attr(,"pH scale")
[1] "free"
```

```
> names(ae)
```

[1]	"S"	"t"	"p"	"T"
[5]	"Cl"	"I"	"P"	"Pa"
[9]	"d"	"density"	"SumCO2"	"SumNH4"
[13]	"SumH2S"	"SumHNO3"	"SumHNO2"	"SumH3PO4"
[17]	"SumSiOH4"	"SumBOH3"	"SumH2SO4"	"SumHF"
[21]	"Br"	"ClConc"	"Na"	"Mg"
[25]	"Ca"	"K"	"Sr"	"molal2molin"
[29]	"free2tot"	"free2sws"	"tot2free"	"tot2sws"
[33]	"sws2free"	"sws2tot"	"K0_CO2"	"K0_O2"
[37]	"fCO2atm"	"fO2atm"	"CO2_sat"	"O2_sat"
[41]	"K_W"	"K_HSO4"	"K_HF"	"K_CO2"
[45]	"K_HCO3"	"K_BOH3"	"K_NH4"	"K_H2S"
[49]	"K_H3PO4"	"K_H2PO4"	"K_HPO4"	"K_SiOH4"
[53]	"K_SiOOH3"	"K_HNO2"	"K_HNO3"	"K_H2SO4"
[57]	"K_HS"	"Ksp_calcite"	"Ksp_aragonite"	"TA"
[61]	"pH"	"fCO2"	"CO2"	"HCO3"
[65]	"CO3"	"BOH3"	"BOH4"	"OH"
[69]	"H3PO4"	"H2PO4"	"HPO4"	"PO4"
[73]	"SiOH4"	"SiOOH3"	"SiO2OH2"	"H2S"
[77]	"HS"	"S2min"	"NH4"	"NH3"
[81]	"H2SO4"	"HSO4"	"SO4"	"HF"
[85]	"F"	"HNO3"	"NO3"	"HNO2"
[89]	"NO2"	"omega_calcite"	"omega_aragonite"	

As seen above, a full speciation is calculated along with the pH or total alkalinity respectively. If only pH or total alkalinity is needed, the calculation of the full speciation can be toggled off. Furthermore, the flag `skeleton` also works for a full system.

```
> ae <- aquaenv(S, t, p, SumCO2 = SumCO2, pH = pH, speciation = FALSE)
> names(ae)
```

[1]	"S"	"t"	"p"	"T"
[5]	"Cl"	"I"	"P"	"Pa"
[9]	"d"	"density"	"SumCO2"	"SumNH4"
[13]	"SumH2S"	"SumHNO3"	"SumHNO2"	"SumH3PO4"
[17]	"SumSiOH4"	"SumBOH3"	"SumH2SO4"	"SumHF"
[21]	"Br"	"ClConc"	"Na"	"Mg"
[25]	"Ca"	"K"	"Sr"	"molal2molin"
[29]	"free2tot"	"free2sws"	"tot2free"	"tot2sws"
[33]	"sws2free"	"sws2tot"	"K0_CO2"	"K0_O2"
[37]	"fCO2atm"	"fO2atm"	"CO2_sat"	"O2_sat"
[41]	"K_W"	"K_HSO4"	"K_HF"	"K_CO2"
[45]	"K_HCO3"	"K_BOH3"	"K_NH4"	"K_H2S"
[49]	"K_H3PO4"	"K_H2PO4"	"K_HPO4"	"K_SiOH4"
[53]	"K_SiOOH3"	"K_HNO2"	"K_HNO3"	"K_H2SO4"
[57]	"K_HS"	"Ksp_calcite"	"Ksp_aragonite"	"TA"
[61]	"pH"	"fCO2"	"CO2"	

```
> ae <- aquaenv(S, t, p, SumCO2 = SumCO2, pH = pH, speciation = FALSE,
+   skeleton = TRUE)
> names(ae)
```

```
[1] "S"      "t"      "p"      "T"      "Cl"     "I"
[7] "P"      "Pa"     "d"      "density" "SumCO2" "SumNH4"
[13] "SumH2S" "SumHNO3" "SumHNO2" "SumH3PO4" "SumSiOH4" "SumBOH3"
[19] "SumH2SO4" "SumHF" "K_W" "K_HSO4" "K_HF" "K_CO2"
[25] "K_HCO3" "K_BOH3" "K_NH4" "K_H2S" "K_H3PO4" "K_H2PO4"
[31] "K_HPO4" "K_SiOH4" "K_SiOOH3" "K_HNO2" "K_HNO3" "K_H2SO4"
[37] "K_HS" "TA" "pH" "fCO2" "CO2"
```

Furtermore all the quantities needed for the explicit pH modelling approaches as given in Hofmann *et al.* (2008a) and Hofmann *et al.* (2008b) can be calculated by setting the flag `dsa` to TRUE. The Revelle factor can be calculated using the flag `revelle`.

```
> ae <- aquaenv(S, t, p, SumCO2 = SumCO2, fCO2 = fCO2, dsa = TRUE,
+   reveille = TRUE)
> ae$dTAdH
```

```
[1] -17222.15
attr("unit")
[1] "(mol-TA/kg-soln)/(mol-H/kg-soln)"
attr("pH scale")
[1] "free"
```

```
> ae$revelle
```

```
[1] 13.71635
```

If an ambivalent situation is created because the user enters too much information, an error message is displayed

```
> ae <- aquaenv(S, t, p, SumCO2 = SumCO2, CO2 = CO2, fCO2 = fCO2)

[1] "Error! Overdetermined system: entered fCO2: 0.0005272996 , calculated fCO2: 0.00052729949660769"
[1] "Please enter only one of: pH, TA, CO2, or fCO2."

> ae <- aquaenv(S, t, p, SumCO2 = SumCO2, pH = pH, TA = TA)

[1] "Error! Overdetermined system: entered TA: 0.002142233 , calculated TA: 0.00214224262579018"
[1] "Please enter only one of: pH, TA, CO2, or fCO2."

> ae <- aquaenv(S, t, p, SumCO2 = SumCO2, pH = pH, CO2 = CO2)

[1] "Error! Overdetermined system: entered pH: 8 , calculated pH: 7.9999746670567"
[1] "Please enter only one of: pH, TA, CO2, or fCO2."

> ae <- aquaenv(S, t, p, SumCO2 = SumCO2, pH = pH, fCO2 = fCO2)
```



```
[1] "Error! Overdetermined system: entered pH: 8 , calculated pH: 7.99997458567836"
[1] "Please enter only one of: pH, TA, CO2, or fCO2."

> ae <- aquaenv(S, t, p, SumCO2 = SumCO2, TA = TA, CO2 = CO2)

[1] "Error! Overdetermined system: entered TA: 0.002142233 , calculated TA: 0.00214223257912512"
[1] "Please enter only one of: pH, TA, CO2, or fCO2."

> ae <- aquaenv(S, t, p, SumCO2 = SumCO2, TA = TA, fCO2 = fCO2)

[1] "Error! Overdetermined system: entered TA: 0.002142233 , calculated TA: 0.00214223254685228"
[1] "Please enter only one of: pH, TA, CO2, or fCO2."
```

3.1.4 Calculating $[\sum \text{CO}_2]$

$[\sum \text{CO}_2]$ can be calculated by giving a constant pair of either pH and CO₂, pH and fCO₂, pH and TA, TA and CO₂, or TA and fCO₂

```
> fCO2 <- 0.0006943363
> CO2 <- 2.674693e-05
> pH <- 7.884892
> TA <- 0.0021
> S <- 30
> t <- 15
> p <- 10
> ae <- aquaenv(S, t, p, SumCO2 = NULL, pH = pH, CO2 = CO2)
> ae$SumCO2

[1] 0.00200012
attr("unit")
[1] "mol/kg-soln"

> ae <- aquaenv(S, t, p, SumCO2 = NULL, pH = pH, fCO2 = fCO2)
> ae$SumCO2

[1] 0.002000120
attr("unit")
[1] "mol/kg-soln"

> ae <- aquaenv(S, t, p, SumCO2 = NULL, pH = pH, TA = TA)
> ae$SumCO2

[1] 0.001999992
attr("unit")
[1] "mol/kg-soln"

> ae <- aquaenv(S, t, p, SumCO2 = NULL, TA = TA, CO2 = CO2)
> ae$SumCO2
```

```
[1] 0.002
attr("unit")
[1] "mol/kg-soln"
```

```
> ae <- aquaenv(S, t, p, SumCO2 = NULL, TA = TA, fCO2 = fCO2)
> ae$SumCO2
```

```
[1] 0.002
attr("unit")
[1] "mol/kg-soln"
```

3.1.5 Cloning an object of class *aquaenv*

It is possible to clone an object of class *aquaenv*, either 1 to 1 or with different pH, TA, or K_{CO2}

```
> S <- 30
> t <- 15
> SumCO2 <- 0.002
> TA <- 0.00214
> ae <- aquaenv(S, t, SumCO2 = SumCO2, TA = TA)
> ae$pH
```

```
[1] 7.998381
attr("pH scale")
[1] "free"
```

```
> ae1 <- aquaenv(ae = ae)
> ae1$pH
```

```
[1] 7.998381
attr("pH scale")
[1] "free"
```

```
> pH <- 9
> ae2 <- aquaenv(ae = ae, pH = pH)
> ae2$TA
```

```
[1] 0.002982756
attr("unit")
[1] "mol/kg-soln"
```

```
> TA <- 0.002
> ae3 <- aquaenv(ae = ae, TA = TA)
> ae3$pH
```

```
[1] 7.548175
attr(,"pH scale")
[1] "free"

> K_CO2 <- 1e-06
> ae4 <- aquaenv(ae = ae, k_co2 = K_CO2)
> ae4$pH
```

```
[1] 7.998381
attr(,"pH scale")
[1] "free"
```

Note that K_{CO2} as an input variable is in lower cases!

3.1.6 Preparing input variables

Input variables for the function `aquaenv` need to be in mol/kg-solution and on the free pH scale. Data in other concentration units or pH scales can be converted using the function `convert`.

```
> S <- 10
> t <- 15
> pH_NBS <- 8.142777
> SumCO2molar <- 0.002016803
> pH_free <- convert(pH_NBS, "pHscale", "nbs2free", S = S, t = t)
> SumCO2molin <- convert(SumCO2molar, "conc", "molar2molin", S = S,
+   t = t)
> ae <- aquaenv(S, t, SumCO2 = SumCO2molin, pH = pH_free)
> ae$pH

[1] 8.004386
attr(,"pH scale")
[1] "free"

> ae$SumCO2

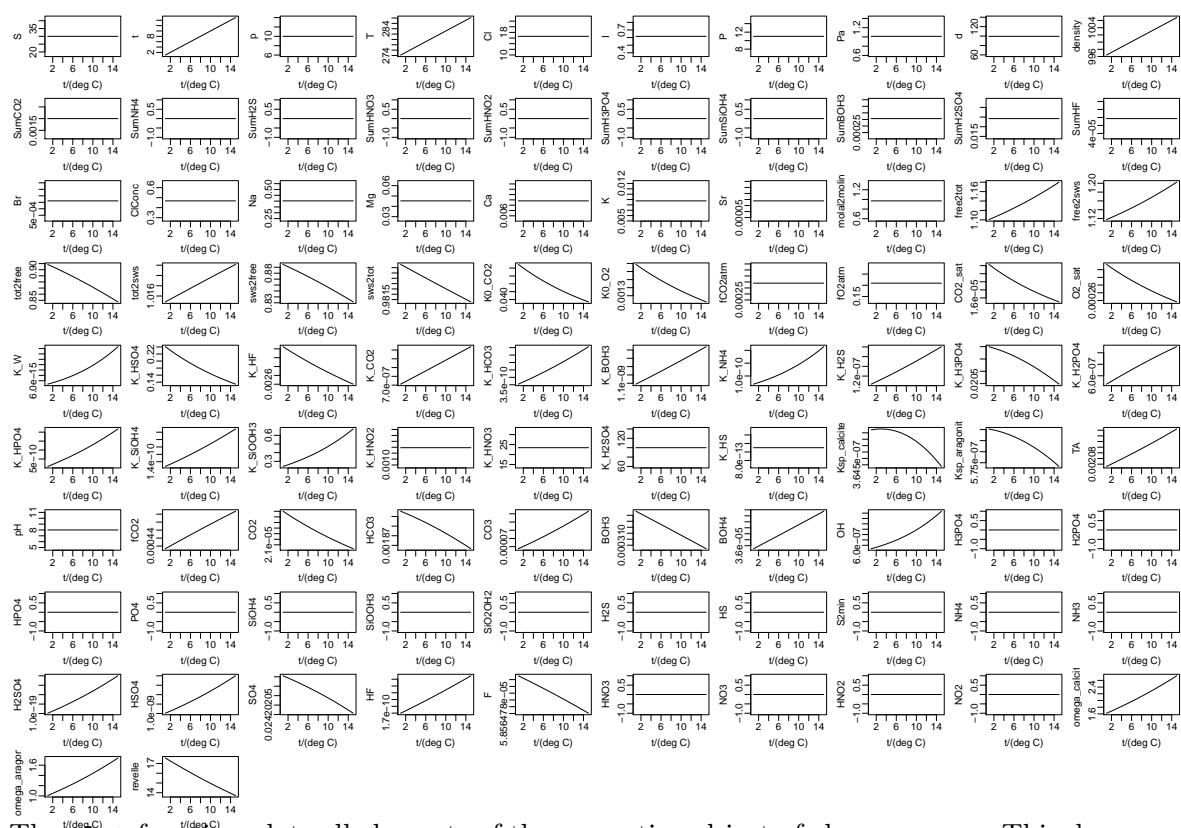
[1] 0.002003213
attr(,"unit")
[1] "mol/kg-soln"
```

3.1.7 Vectors as input variables

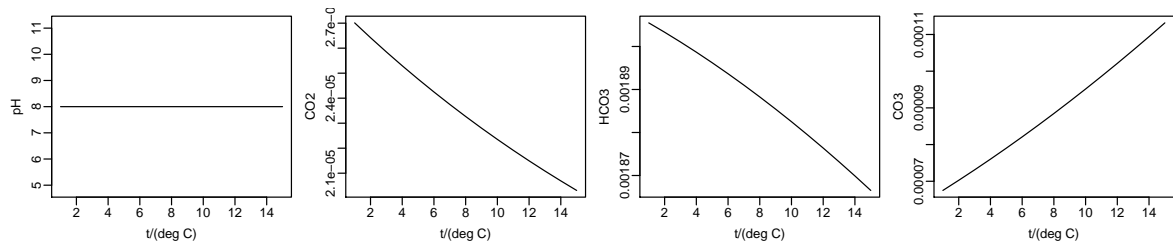
One of the input variables for the function `aquaenv` may be a vector. All the other input variables are then assumed to be constant. The elements of the resulting two dimensional object of class `aquaenv` are then vectors containing the elements as a function of the input variable for which a vector is given.

[1] 17.55733 17.25138 16.94892 16.65047 16.35644 16.06720 15.78307 15.50430
[9] 15.23112 14.96368 14.70212 14.44654 14.19698 13.95350 13.71609

```
> plot(ae, xval = t, xlab = "t/(deg C)", newdevice = FALSE)
```



```
> plot(ae, xval = t, xlab = "t/(deg C)", what = c("pH", "CO2",  
+ "HCO3", "CO3"), newdevice = FALSE, mfrow = c(1, 4))
```



The following chunks of example code show other possible definitions of objects of class *aquaenv* with vectors as input variables.

```
> ae <- aquaenv(S = 20:30, t = 15, p = 10, SumCO2 = SumCO2, pH = pH,
+   dsa = TRUE)
> plot(ae, xval = 20:30, xlab = "S")

> ae <- aquaenv(S = 30, t = 15, p = seq(1, 1000, 100), SumCO2 = SumCO2,
+   pH = pH, revelle = TRUE)
> plot(ae, xval = seq(1, 1000, 100), xlab = "gauge p/bar")
> ae <- aquaenv(S = 30, t = 15, d = seq(1, 10000, 1000), SumCO2 = SumCO2,
+   pH = pH, revelle = TRUE)
> plot(ae, xval = seq(1, 10000, 1000), xlab = "depth/m")

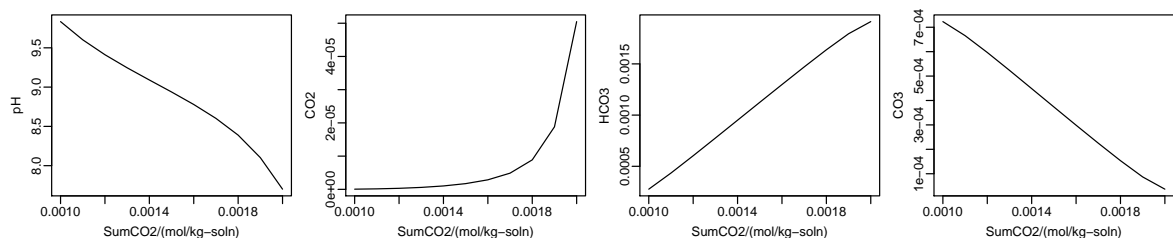
> ae <- aquaenv(S = 30, t = 1:15, p = 10, SumCO2 = SumCO2, TA = 0.0023)
> plot(ae, xval = 1:15, xlab = "t/(deg C)")

> ae <- aquaenv(S = 20:30, t = 15, p = 10, SumCO2 = SumCO2, TA = TA)
> plot(ae, xval = 20:30, xlab = "S")

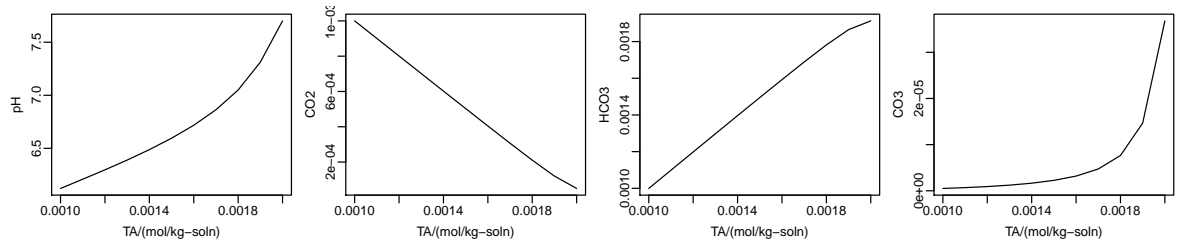
> ae <- aquaenv(S = 30, t = 15, d = seq(1, 1000, 200), SumCO2 = SumCO2,
+   TA = TA, revelle = TRUE, dsa = TRUE)
> plot(ae, xval = seq(1, 1000, 200), xlab = "depth/m")
```

Interesting to note is that also, e.g., SumCO₂, TA, pH and SumNH₄ can be vectors

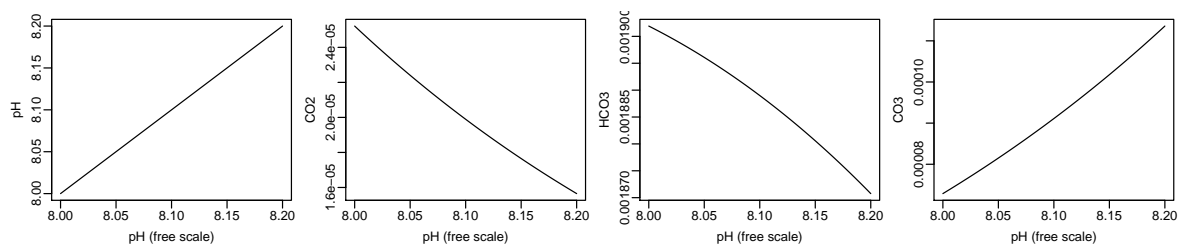
```
> ae <- aquaenv(20, 10, SumCO2 = seq(0.001, 0.002, 1e-04), TA = 0.002)
> plot(ae, xval = ae$SumCO2, xlab = "SumCO2/(mol/kg-soln)", what = c("pH",
+   "CO2", "HCO3", "CO3"), newdevice = FALSE, mfrow = c(1, 4))
```



```
> ae <- aquaenv(20, 10, SumCO2 = 0.002, TA = seq(0.001, 0.002,
+   1e-04))
> plot(ae, xval = ae$TA, xlab = "TA/(mol/kg-soln)", what = c("pH",
+   "CO2", "HCO3", "CO3"), newdevice = FALSE, mfrow = c(1, 4))
```



```
> ae <- aquaenv(20, 10, SumCO2 = 0.002, pH = seq(8, 8.2, 0.001))
> plot(ae, xval = ae$pH, xlab = "pH (free scale)", what = c("pH",
+ "CO2", "HCO3", "CO3"), newdevice = FALSE, mfrow = c(1, 4))
```



```
> ae <- aquaenv(20, 10, SumCO2 = 0.002, SumNH4 = seq(1e-04, 2e-04,
+ 1e-05), pH = 8)
> ae$NH3
```

```
[1] 1.534559e-06 1.688014e-06 1.841470e-06 1.994926e-06 2.148382e-06
[6] 2.301838e-06 2.455294e-06 2.608749e-06 2.762205e-06 2.915661e-06
[11] 3.069117e-06
attr("unit")
[1] "mol/kg-soln"
```

3.1.8 Calculating $\sum \text{CO}_2$ from input vectors

The functionality of calculating SumCO2 can also be used together with vectors as input variables.

```
> ae <- aquaenv(S = 30, t = 11:15, SumCO2 = NULL, pH = pH, CO2 = CO2,
+ revelle = TRUE, dsa = TRUE)
> ae$SumCO2
```

```
[1] 0.002412872 0.002460282 0.002507956 0.002555877 0.002604026
attr("unit")
[1] "mol/kg-soln"
```

Two further examples

```
> ae <- aquaenv(S = 20:30, t = 15, SumCO2 = NULL, pH = pH, fCO2 = fCO2)
> plot(ae, xval = 20:30, xlab = "S")
```

```
> ae <- aquaenv(S = 30, t = 15, d = seq(1, 1000, 100), SumCO2 = NULL,
+             pH = pH, TA = TA)
> plot(ae, xval = seq(1, 1000, 100), xlab = "depth/m")
```

3.1.9 Conversion from and to a dataframe

Objects of class *aquaenv* can be converted to an R *data.frame* to further post-process them with standard R means. Similarly, R *data.frames* can be converted to objects of class *aquaenv* to use the plotting facilities that exist for objects of class *aquaenv*. This can be helpful for plotting output of a dynamic model run, e.g. from R package **deSolve**, and will be shown later in this document.

```
> aedataframe <- as.data.frame(ae)
> aetest <- aquaenv(ae = aedataframe, from.data.frame = TRUE)
```

3.1.10 Converting elements in an object of class *aquaenv*

Elements of an object of class *aquaenv* are calculated in, e.g., the concentration unit mol/kg-solution (molality). The function **convert** can be used to convert all elements in an object of class *aquaenv* that share a common attribute, e.g. the unit.

```
> ae <- aquaenv(30, 10)
> ae$SumBOH3

[1] 0.0003563636
attr("unit")
[1] "mol/kg-soln"

> ae <- convert(ae, "mol/kg-soln", "umol/kg-H2O", 1e+06/ae$molal2molin,
+             "unit")
> ae$SumBOH3

[1] 367.442
attr("unit")
[1] "umol/kg-H2O"
```

3.1.11 Quantities needed for explicit pH modelling

As already mentioned above, the quantities needed for the explicit pH modelling approach (direct substitution approach - DSA) as presented by Hofmann *et al.* (2008a) can be calculated with the function **aquaenv** by setting the flag **dsa**.

```
> ae <- aquaenv(S = 30, t = 15, d = 10, SumCO2 = 0.002, pH = 8,
+             dsa = TRUE, revelle = TRUE)
```

This command calculated the buffer factor and the partial derivatives of [TA] with respect to other summed quantities referred to in Hofmann *et al.* (2008a)

```

> ae$dTAdH

[1] -17142.02
attr(,"unit")
[1] "(mol-TA/kg-soln)/(mol-H/kg-soln)"
attr(,"pH scale")
[1] "free"

> ae$dTAdSumCO2

[1] 1.045937
attr(,"unit")
[1] "(mol-TA/kg-soln)/(mol-SumCO2/kg-soln)"

```

as well the sums partial derivatives of [TA] with respect to the equilibrium constants (K^* 's) multiplied with the partial derivatives of the respective equilibrium constant with one of their variables (i.e., S, T, d, SumH2SO4, od SumHF) as introduced in [Hofmann *et al.* \(2009\)](#).

```

> ae$dTAdKdKdS

[1] 3.986489e-06
attr(,"unit")
[1] "(mol-TA/kg-soln)/\"psu\""

> ae$dTAdKdKdSumH2SO4

[1] -0.001086872
attr(,"unit")
[1] "(mol-TA/kg-soln)/(mol-SumH2SO4/kg-soln)"

```

Furthermore, the ionization fractions used for the pH dependent fractional stoichiometric pH modelling approach described in [Hofmann *et al.* \(2008b\)](#) are calculated as well

```

> ae$c1

[1] 0.01025967

```

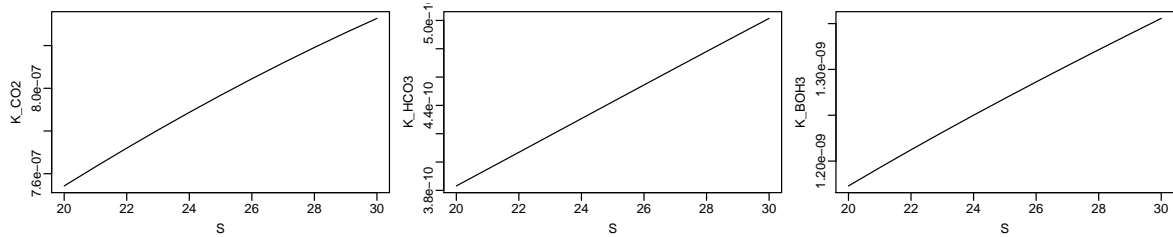
3.2 The `plot.aquaenv` function

In the previous sections, the `plot` function has been introduced. What actually is called if the first element of the arguments list of `plot` is an object pf class *aquaenv* is the function `plot.aquaenv`. This is a multifunctional tool to visualize information contained in an object of class *aquaenv*. For the convenience of the users, `plot.aquaenv` combines the call of standard R plotting functions and the previous call of the function `par` to set parameters like `mfrow`, `mar`, etc. as well as the opening of a plotting device with a certain size. As already shown above, setting the flag `newdevice` to `FALSE` suppresses the opening of a new plotting

device (this feature is needed here to create a plot that will be woven into the L^AT_EX document by Sweave).

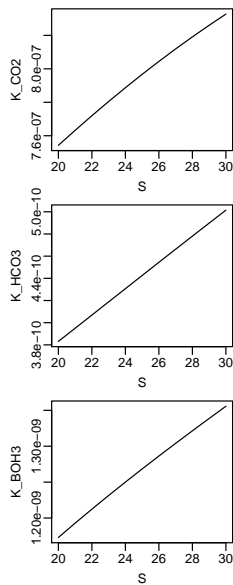
For example

```
> ae <- aquaenv(20:30, 10)
> plot(ae, xval = 20:30, xlab = "S", what = c("K_CO2", "K_HCO3",
+       "K_BOH3"), size = c(10, 2), mfrow = c(1, 3), newdevice = FALSE)
```



and

```
> plot(ae, xval = 20:30, xlab = "S", what = c("K_CO2", "K_HCO3",
+       "K_BOH3"), size = c(2, 10), mfrow = c(3, 1), newdevice = FALSE)
```

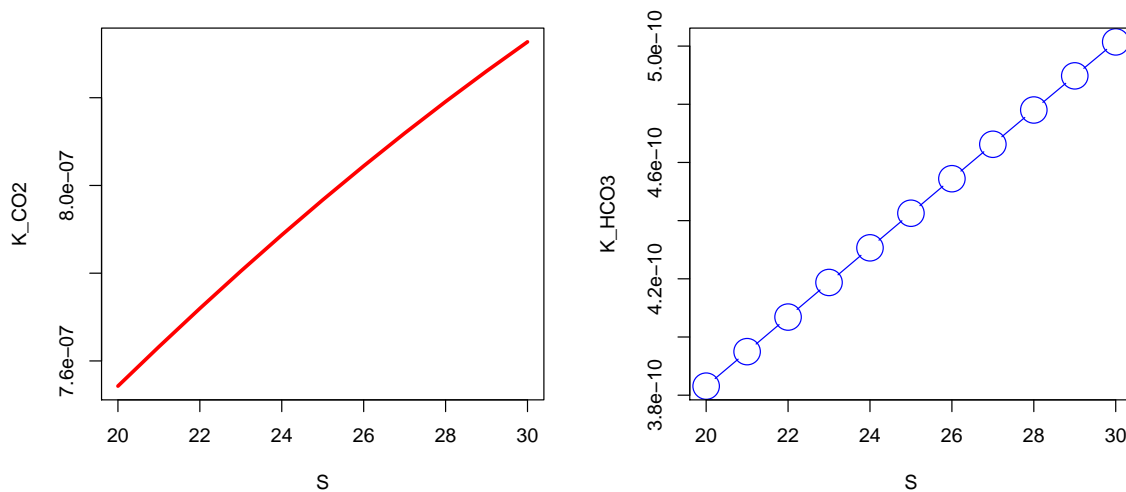


Furthermore the parameter `device` can be specified which allows the user to write the plots to .eps and .pdf files. The parameter `filename` can be used to specify a filename other than the default filename “aquaenv”.

```
> ae <- aquaenv(20:30, 10)
> plot(ae, xval = 20:30, xlab = "S", what = c("K_CO2", "K_HCO3",
+       "K_BOH3"), size = c(10, 2), mfrow = c(1, 3), device = "pdf",
+       filename = "test")
> plot(ae, xval = 20:30, xlab = "S", what = c("K_CO2", "K_HCO3",
+       "K_BOH3"), size = c(2, 10), mfrow = c(3, 1), device = "eps",
+       filename = "test")
```

These features make the function `plot.aquaenv` different from standard R plotting functions. However, if the flags `newdevice` and `setpar` are set to `FALSE`, `plot.aquaenv` behaves like a “normal” R plotting function

```
> par(mfrow = c(1, 2))
> plot(ae, xval = 20:30, xlab = "S", what = "K_CO2", lwd = 3, col = "red",
+      newdevice = FALSE, setpar = FALSE)
> plot(ae, xval = 20:30, xlab = "S", what = "K_HCO3", cex = 3,
+      type = "b", col = "blue", newdevice = FALSE, setpar = FALSE)
```



Furthermore, the function `plot.aquaenv` can be used to create “cumulative” plots and “Bjerrum” plots. This will be explained in some of the following sections.

3.3 Using objects of class *aquaenv* in dynamic models

3.3.1 Ordinary dynamic models

It is convenient to use objects of class *aquaenv* in a dynamic model, e.g. solved using the R package **deSolve**. This can be illustrated with an example. (For information about how to set up a dynamic model with **deSolve**, consult the documentation of **deSolve**).

A list of parameters is specified

```
> parameters <- list(
+   S      = 25      , # psu
+   t_min  = 5       , # degrees C
+   t_max  = 25      , # degrees C
+
+   k      = 0.4      , # 1/d                proportionality factor for air-water exchange
+   rOx    = 0.0000003 , # mol-N/(kg*d)    maximal rate of oxic mineralisation
+   rNitri = 0.0000002 , # mol-N/(kg*d)    maximal rate of nitrification
+   rPP    = 0.000006  , # mol-N/(kg*d)    maximal rate of primary production
+ )
```

```

+ ksDINPP = 0.000001 , # mol-N/kg
+ ksNH4PP = 0.000001 , # mol-N/kg
+
+ D = 0.1 , # 1/d (dispersive) transport coefficient
+
+ O2_io = 0.000296 , # mol/kg-soln
+ NO3_io = 0.000035 , # mol/kg-soln
+ SumNH4_io = 0.000008 , # mol/kg-soln
+ SumCO2_io = 0.002320 , # mol/kg-soln
+ TA_io = 0.002435 , # mol/kg-soln
+
+ C_Nratio = 8 , # mol C/mol N C:N ratio of organic matter
+
+ a = 30 , # timestep from which PP begins
+ b = 50 , # timestep where PP shuts off again
+
+ modeltime = 100 # duration of the model
+ )

```

A model function is defined which will be executed every timestep of the numerical integration. An object of class *aquaenv* is created in each timestep, some of its elements are used to calculate kinetic rate expressions and the whole object is returned as output.

```

> boxmodel <- function(timestep, currentstate, parameters)
+ {
+   with (
+     as.list(c(currentstate,parameters)),
+     {
+       t <- c(seq(t_min, t_max, (t_max-t_min)/(modeltime/2)),
+             seq(t_max, t_min, -(t_max-t_min)/(modeltime/2)))[round(timestep)+1]]
+
+       ae <- aquaenv(S=S, t=t, SumCO2=SumCO2, SumNH4=SumNH4, TA=TA)
+
+       ECO2 <- k * (ae$CO2_sat - ae$CO2)
+       EO2 <- k * (ae$O2_sat - O2)
+
+       TO2 <- D*(O2_io - O2)
+       TNO3 <- D*(NO3_io - NO3)
+       TSumNH4 <- D*(SumNH4_io - SumNH4)
+       TTA <- D*(TA_io - TA)
+       TSumCO2 <- D*(SumCO2_io - SumCO2)
+
+       RNit <- rNitri
+
+       ROx <- rOx
+       ROxCarbon <- ROx * C_Nratio
+
+       pNH4PP <- 0
+       RPP <- 0
+
+       if ((timestep > a) && (timestep < b))
+       {
+         RPP <- rPP * ((SumNH4+NO3)/(ksDINPP + (SumNH4+NO3)))
+         pNH4PP <- 1 - (ksNH4PP/(ksNH4PP + SumNH4))
+       }
+       else
+       {
+         RPP <- 0
+       }
+
+       RPPCarbon <- RPP * C_Nratio
+
+       dO2 <- TO2 + EO2 - ROxCarbon - 2*RNit + (2-2*pNH4PP)*RPP + RPPCarbon
+       dNO3 <- TNO3 + RNit - (1-pNH4PP)*RPP
+
+       dSumCO2 <- TSumCO2 + ECO2 + ROxCarbon - RPPCarbon
+

```

```

+       dSumNH4 <- TSumNH4 + ROx - RNit - pNH4PP*RPP
+
+       dTA      <- TTA      + ROx - 2*RNit -(2*pNH4PP-1)*RPP
+
+       ratesofchanges <- c(dO2, dNO3, dSumNH4, dSumCO2, dTA)
+       transport      <- c(TO2=TO2, TNO3=TNO3, TSumNH4=TSumNH4, TTA=TTA, TSumCO2=TSumCO2)
+       airseaexchange <- c(ECO2=ECO2, EO2=EO2)
+
+       return(list(ratesofchanges, list(transport, airseaexchange, ae)))
+     }
+   )
+ }

```

The model is solved.

```

> with (as.list(parameters),
+   {
+     initialstate <- c(O2=O2_io, NO3=NO3_io, SumNH4=SumNH4_io, SumCO2=SumCO2_io, TA=TA_io)
+     times        <- c(0:modeltime)
+     output       <- as.data.frame(vode(initialstate,times,boxmodel,parameters, hmax=1))[-1,]
+   })

```

and the output can be plotted in the same way as a two dimensional object of class *aquaenv* by converting it to an object of class *aquaenv* using the `from.data.frame` flag of the function `aquaenv`. (Please note that due to runtime constraints of the vignette for this package, this model is not run during construction of the vignette and, therefore, no resulting plot of the following code chunk is included. The user can generate the plot by extracting and executing the relevant code-chunks.)

```

> plot(aquaenv(ae = output, from.data.frame = TRUE), xval = output$time,
+   xlab = "time/d", mfrow = c(11, 10), newdevice = FALSE)

```

3.3.2 Models using the explicit pH modelling approach

3.3.2.1 In one single model

Since an object of class *aquaenv* can contain all quantities necessary to employ the explicit pH modelling approaches as introduced by Hofmann *et al.* (2008a, 2009, 2008b), they can be readily used in an explicit pH model.

As an example, we give a model that calculates the pH in the “classical” way in every timestep using *aquaenv*, also employs the explicit pH modelling approach (direct substitution approach - DSA) given in Hofmann *et al.* (2008a) and additionally employs fractional stoichiometry as given in Hofmann *et al.* (2008b). The pH evolution is thus calculated in three different ways which allows comparing the three values for consistency.

Again, a list of parameters is defined

```

> parameters <- list(
+   S      = 25      , # psu
+   t      = 15      , # degrees C
+
+   k      = 0.4      , # 1/d      proportionality factor for air-water exchange

```

```

+   rOx      = 0.0000003 , # mol-N/(kg*d)  maximal rate of oxic mineralisation
+   rNitri   = 0.0000002 , # mol-N/(kg*d)  maximal rate of nitrification
+   rPP      = 0.0000006 , # mol-N/(kg*d)  maximal rate of primary production
+
+   ksSumNH4 = 0.000001  , # mol-N/kg
+
+   D        = 0.1      , # 1/d              (dispersive) transport coefficient
+
+   O2_io     = 0.000296 , # mol/kg-soln
+   NO3_io    = 0.000035 , # mol/kg-soln
+   SumNH4_io = 0.000008 , # mol/kg-soln
+   SumCO2_io = 0.002320 , # mol/kg-soln
+   TA_io     = 0.002435 , # mol/kg-soln
+
+   C_Nratio  = 8        , # mol C/mol N    C:N ratio of organic matter
+
+   a         = 30       , # timestep from which PP begins
+   b         = 50       , # timestep where PP shuts off again
+
+   modeltime = 100      # duration of the model
+ )

```

And a model function is defined. Again, an object of class *aquaenv* is created in each timestep and respective elements are used.

```

> boxmodel <- function(timestep, currentstate, parameters)
+ {
+   with (
+     as.list(c(currentstate,parameters)),
+     {
+       # the "classical" implicit pH calculation method is applied in aquaenv
+       ae <- aquaenv(S=S, t=t, SumCO2=SumCO2, SumNH4=SumNH4, TA=TA, dsa=TRUE)
+
+       ECO2 <- k * (ae$CO2_sat - ae$CO2)
+       EO2  <- k * (ae$O2_sat - O2)
+
+       RNit <- rNitri
+       ROx  <- rOx
+
+       if ((timestep > a) && (timestep < b))
+       {
+         RPP <- rPP * (SumNH4/(ksSumNH4 + SumNH4))
+       }
+       else
+       {
+         RPP <- 0
+       }
+
+       dO2 <- EO2 - C_Nratio*ROx - 2*RNit + C_Nratio*RPP
+       dNO3 <- RNit
+
+       dSumCO2 <- ECO2 + C_Nratio*ROx - C_Nratio*RPP
+       dSumNH4 <- ROx - RNit - RPP
+
+       dTA <- ROx - 2*RNit - RPP
+
+       # The DSA pH
+       dH <- (dTA - (dSumCO2*ae$dTAdSumCO2 + dSumNH4*ae$dTAdSumNH4))/ae$dTAdH
+       DSApH <- -log10(H)
+
+       # The DSA pH using pH dependent fractional stoichiometry (= using partitioning coefficients)
+       rhoHECO2 <- ae$c2 + 2*ae$c3
+       rhoHRNit <- 1 + ae$n1
+       rhoHROx <- C_Nratio * (ae$c2 + 2*ae$c3) - ae$n1
+       rhoHRPP <- -(C_Nratio * (ae$c2 + 2*ae$c3)) + ae$n1
+     }
+   )
+ }

```

```

+
+     dH_ECO2 <- rhoHECO2*ECO2/(-ae$dTAdH)
+     dH_RNIt <- rhoHRNIt*RNIt/(-ae$dTAdH)
+     dH_ROx <- rhoHROx*ROx /(-ae$dTAdH)
+     dH_RPP <- rhoHRPP*RPP /(-ae$dTAdH)
+
+     dH_stoich <- dH_ECO2 + dH_RNIt + dH_ROx + dH_RPP
+     DSAstoichpH <- -log10(H_stoich)
+
+     ratesofchanges <- c(dO2, dNO3, dSumNH4, dSumCO2, dTA, dH, dH_stoich)
+     processrates <- c(ECO2=ECO2, EO2=EO2, RNIt=RNIt, ROx=ROx, RPP=RPP)
+     DSA <- c(DSApH=DSApH, rhoHECO2=rhoHECO2, rhoHRNIt=rhoHRNIt, rhoHROx=rhoHROx,
+              rhoHRPP=rhoHRPP, dH_ECO2=dH_ECO2, dH_RNIt=dH_RNIt, dH_ROx=dH_ROx,
+              dH_RPP=dH_RPP, DSAstoichpH=DSAstoichpH)
+
+     return(list(ratesofchanges, list(processrates, DSA, ae)))
+   }
+ }
+ }

```

The model is solved

```

> with (as.list(parameters),
+   {
+     H_init <- 10^(-(aquaenv(S=S, t=t, SumCO2=SumCO2_io, SumNH4=SumNH4_io, TA=TA_io,
+                           speciation=FALSE)$pH))
+     initialstate <- c(O2=O2_io, NO3=NO3_io, SumNH4=SumNH4_io, SumCO2=SumCO2_io, TA=TA_io, H=H_init,
+                       H_stoich=H_init)
+     times <- c(0:modeltime)
+     output <- as.data.frame(vode(initialstate, times, boxmodel, parameters, hmax=1))[-1,]
+   })

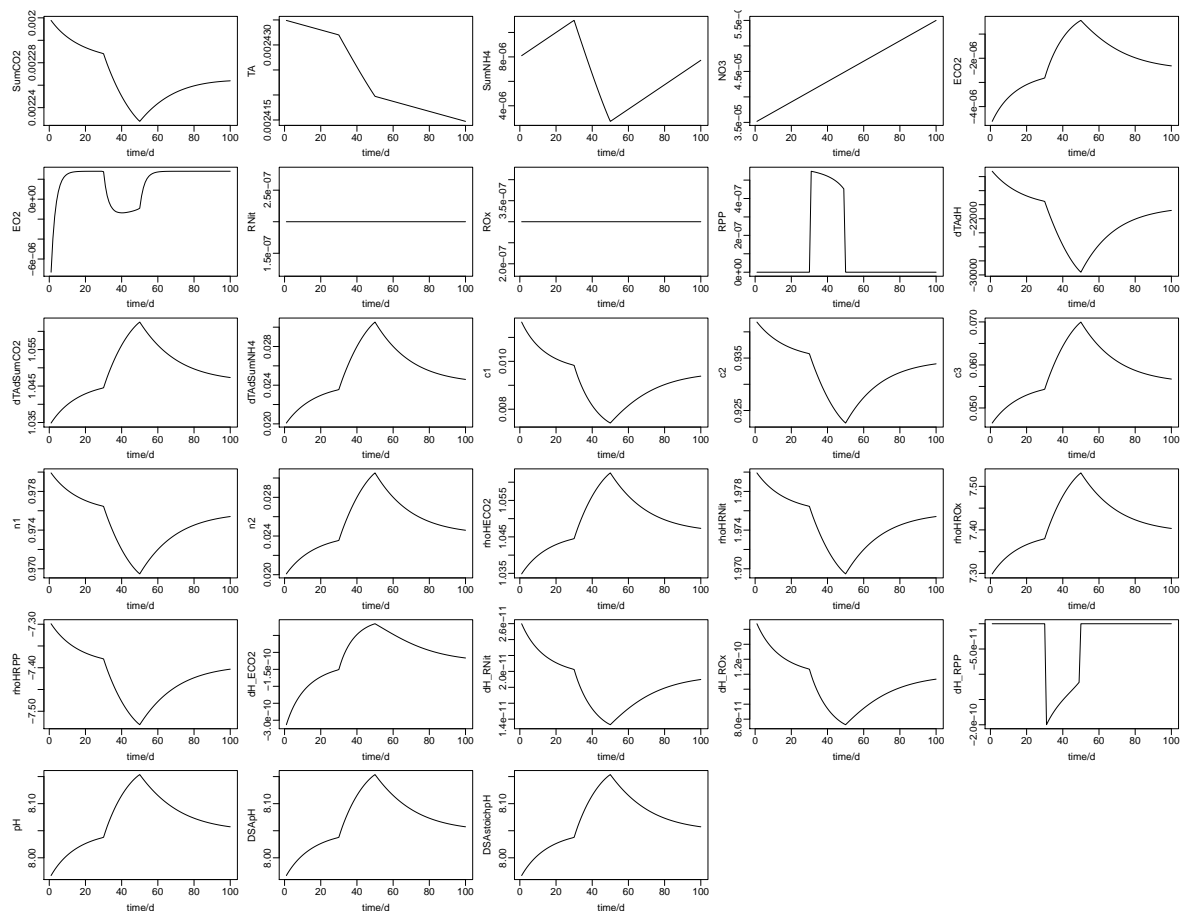
```

and output can be plotted. Again using `plot.aquaenv`. Note that here the parameter `what` is used.

```

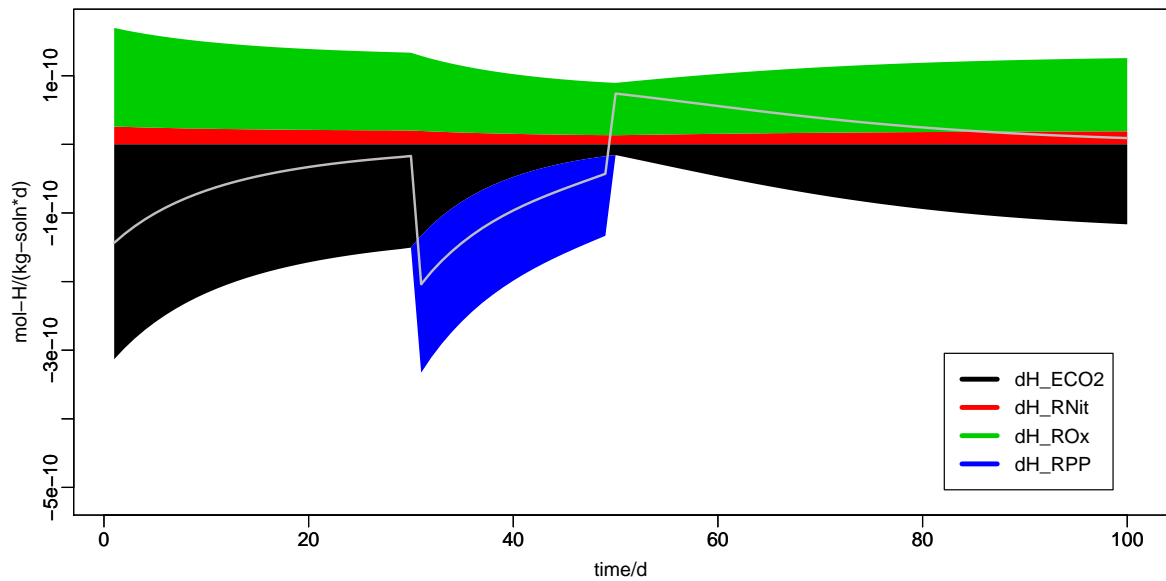
> what <- c("SumCO2", "TA", "SumNH4", "NO3", "ECO2", "EO2", "RNIt",
+          "ROx", "RPP", "dTAdH", "dTAdSumCO2", "dTAdSumNH4", "c1",
+          "c2", "c3", "n1", "n2", "rhoHECO2", "rhoHRNIt", "rhoHROx",
+          "rhoHRPP", "dH_ECO2", "dH_RNIt", "dH_ROx", "dH_RPP", "pH",
+          "DSApH", "DSAstoichpH")
> plot(aquaenv(ae = output, from.data.frame = TRUE), xval = output$time,
+       what = what, xlab = "time/d", mfrow = c(6, 5), size = c(20,
+       13), newdevice = FALSE)

```



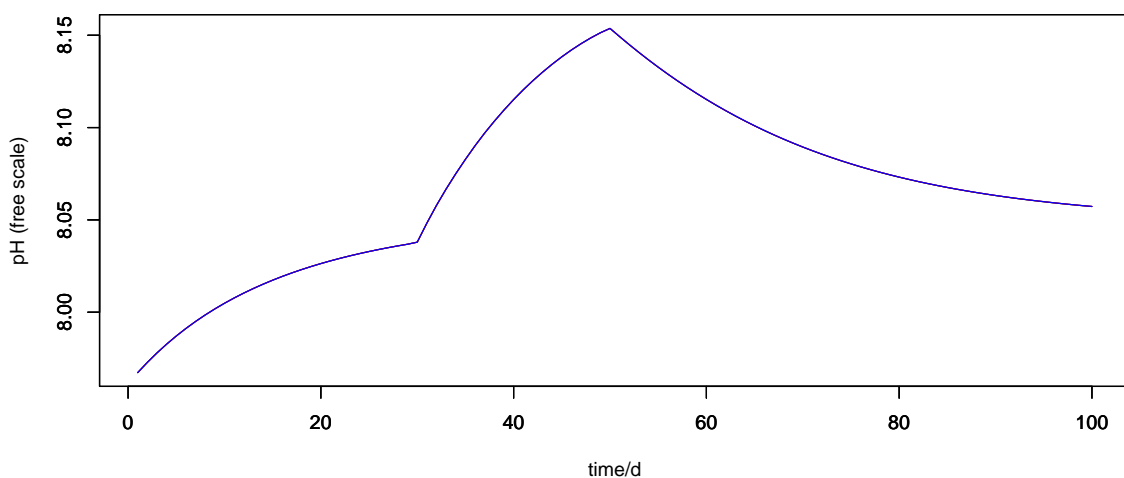
Here, the cumulative plotting functionality of `plot.aquaenv` can be employed as well to visualize the influences of the different kinetically modelled processes on $[H^+]$.

```
> what <- c("dH_ECO2", "dH_RNIt", "dH_ROx", "dH_RPP")
> plot(aquaenv(ae = output, from.data.frame = TRUE), xval = output$time,
+      what = what, xlab = "time/d", size = c(7, 5), ylab = "mol-H/(kg-soln*d)",
+      legendposition = "bottomright", cumulative = TRUE, newdevice = FALSE)
```



Finally, the pH values calculated with the three different methods can be plotted in one single graph to see that they are identical, i.e. the three methods of pH calculation are consistent with each other

```
> ylim <- range(output$DSApH, output$DSAstoichpH, output$pH)
> plot(output$pH, ylim = ylim, type = "l", xlab = "time/d", ylab = "pH (free scale)")
> par(new = TRUE)
> plot(output$DSApH, ylim = ylim, type = "l", col = "red", xlab = "",
+       ylab = "")
> par(new = TRUE)
> plot(output$DSAstoichpH, ylim = ylim, type = "l", col = "blue",
+       xlab = "", ylab = "")
```



3.3.2.2 In three separate models

3.3.2.2.1 The implicit pH modelling approach

A list of parameters

```
> parameters <- list(
+   t           = 15           , # degrees C
+   S           = 35           , # psu
+
+   SumCO2_t0   = 0.002        , # mol/kg-soln (comparable to Wang2005)
+   TA_t0       = 0.0022       , # mol/kg-soln (comparable to Millero1998)
+
+   kc          = 0.5           , # 1/d           proportionality factor for air-water exchange
+   kp          = 0.000001      , # mol/(kg-soln*d) max rate of calcium carbonate precipitation
+   n           = 2.0           , # -           exponent for kinetic rate law of precipitation
+
+   modeltime    = 20           , # d           duration of the model
+   outputsteps  = 100          , #           number of outputsteps
+ )
```

The model function

```
> boxmodel <- function(timestep, currentstate, parameters)
+ {
+   with (
+     as.list(c(currentstate,parameters)),
+     {
+       ae <- aquaenv(S=S, t=t, SumCO2=SumCO2, TA=TA, SumSiOH4=0,
+                     SumBOH3=0, SumH2SO4=0, SumHF=0)
+
+       Rc <- kc * ((ae$CO2_sat) - (ae$CO2))
+       Rp <- kp * (1-ae$omega_calcite)^n
+
+       dSumCO2 <- Rc - Rp
+       dTA <- -2*Rp
+
+       ratesofchanges <- c(dSumCO2, dTA)
+
+       processrates <- c(Rc=Rc, Rp=Rp)
+
+       return(list(ratesofchanges, list(processrates, ae)))
+     }
+   )
+ }
```

Solving the model

```
> with (as.list(parameters),
+   {
+     initialstate <- c(SumCO2=SumCO2_t0, TA=TA_t0)
+     times <- seq(0,modeltime,(modeltime/outputsteps))
+     output <- as.data.frame(vode(initialstate,times,boxmodel,parameters, hmax=1))
+   })
```

Visualizing the output. (Please note that due to runtime constraints of the vignette for this package, this model is not run during construction of the vignette and, therefore, no resulting plot of the following code chunk is included. The user can generate the plot by extracting and executing the relevant code-chunks.)

```
> what <- c("SumCO2", "TA", "Rc", "Rp", "omega_calcite", "pH")
> plot(aquaenv(ae = output, from.data.frame = TRUE), xval = output$time,
+       xlab = "", mfrow = c(2, 3), size = c(12, 5), what = what,
+       newdevice = FALSE)
```

3.3.2.2 The explicit pH modelling approach

A list of parameters

```
> parameters <- list(
+   S           = 35           , # psu
+   t           = 15           , # degrees C
+   SumCO2_t0   = 0.002        , # mol/kg-soln (comparable to Wang2005)
+   TA_t0       = 0.0022       , # mol/kg-soln (comparable to Millero1998)
+   kc          = 0.5          , # 1/d           proportionality factor for air-water exchange
+   kp          = 0.000001     , # mol/(kg-soln*d) max rate of calcium carbonate precipitation
+   n           = 2.0          , # -           exponent for kinetic rate law of precipitation
+   modeltime   = 20           , # d           duration of the model
+   outputsteps = 100          , #           number of outputsteps
+ )
```

The model function

```
> boxmodel <- function(timestep, currentstate, parameters)
+ {
+   with (
+     as.list(c(currentstate,parameters)),
+     {
+       ae <- aquaenv(S=S, t=t, SumCO2=SumCO2, pH=-log10(H), SumSiOH4=0,
+                     SumBOH3=0, SumH2SO4=0, SumHF=0, dsa=TRUE)
+
+       Rc <- kc * ((ae$CO2_sat) - (ae$CO2))
+       Rp <- kp * (1-ae$omega_calcite)^n
+
+       dSumCO2 <- Rc - Rp
+
+       dHRC <- (      -(ae$dTAdSumCO2*Rc      ))/ae$dTAdH
+       dHRp <- (-2*Rp -(ae$dTAdSumCO2*(-Rp)))/ae$dTAdH
+       dH    <- dHRC + dHRp
+
+       ratesofchanges <- c(dSumCO2, dH)
+
+       processrates <- c(Rc=Rc, Rp=Rp)
+       outputvars <- c(dHRC=dHRC, dHRp=dHRp)
+
+       return(list(ratesofchanges, list(processrates, outputvars, ae)))
+     }
+   )
+ }
```

Solving the model

```
> with(as.list(parameters),
+   {
+     aetmp <- aquaenv(S=S, t=t, SumCO2=SumCO2_t0, TA=TA_t0, SumSiOH4=0, SumBOH3=0, SumH2SO4=0, SumHF=0)
+     H_t0 <- 10^(-aetmp$pH)
+
+     initialstate <- c(SumCO2=SumCO2_t0, H=H_t0)
+     times <- seq(0,modeltime,(modeltime/outputsteps))
+     output <- as.data.frame(vode(initialstate,times,boxmodel,parameters, hmax=1))
+   })
```

Visualizing the output. (Please note that due to runtime constraints of the vignette for this package, this model is not run during construction of the vignette and, therefore, no resulting plot of the following two code chunks is included. The user can generate the plot by extracting and executing the relevant code-chunks.)

```
> what <- c("SumCO2", "TA", "Rc", "Rp", "omega_calcite", "pH",
+          "dHRc", "dHRp")
> plot(aquaenv(ae = output, from.data.frame = TRUE), xval = output$time,
+       xlab = "time/d", mfrow = c(3, 3), size = c(15, 10), what = what,
+       newdevice = FALSE)
```

Cumulatively plotting the influences of the two processes on the pH

```
> what <- c("dHRc", "dHRp")
> plot(aquaenv(ae = output, from.data.frame = TRUE), xval = output$time,
+       xlab = "time/d", what = what, ylab = "mol-H/(kg-soln*d)",
+       legendposition = "topright", cumulative = TRUE, newdevice = FALSE)
```

3.3.2.2.3 The fractional stoichiometric approach

A list of parameters

```
> parameters <- list(
+   S      = 35      , # psu
+   t      = 15      , # degrees C
+
+   SumCO2_t0 = 0.002 , # mol/kg-soln (comparable to Wang2005)
+   TA_t0    = 0.0022 , # mol/kg-soln (comparable to Millero1998)
+
+   kc      = 0.5     , # 1/d           proportionality factor for air-water exchange
+   kp      = 0.000001 , # mol/(kg-soln*d)   max rate of calcium carbonate precipitation
+   n       = 2.0     , # -             exponent for kinetic rate law of precipitation
+
+   modeltime = 20     , # d           duration of the model
+   outputsteps = 100  , #             number of outputsteps
+ )
```

The model function

```
> boxmodel <- function(timestep, currentstate, parameters)
+ {
+   with (
+     as.list(c(currentstate, parameters)),
+     {
+       ae <- aquaenv(S=S, t=t, SumCO2=SumCO2, pH=-log10(H), SumSiOH4=0,
+                     SumBOH3=0, SumH2SO4=0, SumHF=0, dsa=TRUE)
+
+       Rc <- kc * ((ae$CO2_sat) - (ae$CO2))
+       Rp <- kp * (1-ae$omega_calcite)^n
+
+       dSumCO2 <- Rc - Rp
+
+       rhoc <- ae$c2 + 2*ae$c3
+       rhop <- 2*ae$c1 + ae$c2
+
+       dHRc <- rhoc*Rc/(-ae$dTAdH)
+       dHRp <- rhop*Rp/(-ae$dTAdH)
```

```

+       dH      <- dHRc + dHRp
+
+       ratesofchanges <- c(dSumCO2, dH)
+
+       processrates  <- c(Rc=Rc, Rp=Rp)
+       outputvars    <- c(dHRc=dHRc, dHRp=dHRp, rhoc=rhoc, rhop=rhop)
+
+       return(list(ratesofchanges, list(processrates, outputvars, ae)))
+     }
+   )
+ }

```

Solving the model

```

> with (as.list(parameters),
+   {
+     aetmp <- aquaenv(S=S, t=t, SumCO2=SumCO2_t0, TA=TA_t0, SumSiOH4=0,
+       SumBOH3=0, SumH2SO4=0, SumHF=0)
+     H_t0 <- 10^(-aetmp$pH)
+
+     initialstate <- c(SumCO2=SumCO2_t0, H=H_t0)
+     times        <- seq(0,modeltime,(modeltime/outputsteps))
+     output       <- as.data.frame(vode(initialstate,times,boxmodel,parameters, hmax=1))
+   })

```

Visualizing the output. (Please note that due to runtime constraints of the vignette for this package, this model is not run during construction of the vignette and, therefore, no resulting plot of the following code chunk is included. The user can generate the plot by extracting and executing the relevant code-chunks.)

```

> what <- c("SumCO2", "TA", "Rc", "Rp", "omega_calcite", "pH",
+   "dHRc", "dHRp", "rhoc", "rhop")
> plot(aquaenv(ae = output, from.data.frame = TRUE), xval = output$time,
+   xlab = "time/d", mfrow = c(3, 4), size = c(15, 10), what = what,
+   newdevice = FALSE)

```

3.4 Titration simulation: the function titration

With the function **titration AquaEnv** provides a powerful tool to simulate titrations. A two dimensional object of class *aquaenv* will be created where the second dimension is the amount of titrant added. For this purpose, three extra elements are added to the *aquaenv* object that will be created:

element	unit	explanation
delta_conc_titrant	mol/kg-solution	the offset in concentration of the titrant that is caused by adding the titrant to the sample
delta_mass_titrant	kg	the amount of mass of titrant solution added
delta_moles_titrant	mol	the amount of moles of titrant added

Each one of this elements is a suitable *xval* for plotting an *aquaenv* object generated by **titration**.

3.4.1 Titration with HCl

The standard titration type is titration with hydrochloric acid (HCl). A simple example will illustrate this best.

An object of class *aquaenv* needs to be created to define the initial conditions of the titration. That is temperature, salinity, depth, the concentrations of all summed quantities and the initial pH (or [TA]).

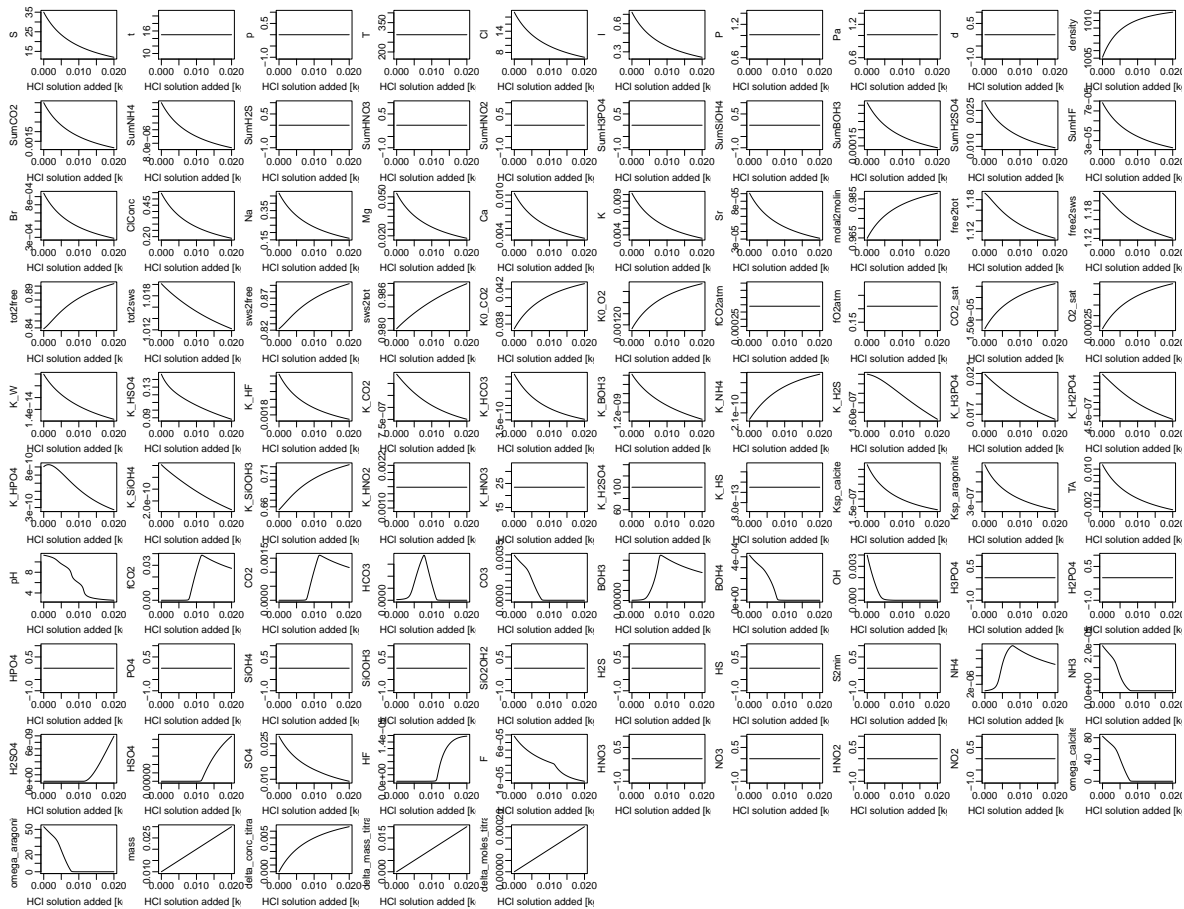
```
> ae_init <- aquaenv(S = 35, t = 15, SumCO2 = 0.0035, SumNH4 = 2e-05,  
+   pH = 11.3)
```

Then `titration` can be run to create the object describing the simulated titration. In this example the titrant is HCl of the relatively low concentration of 0.01 mol/kg-solution. The sample solution amounts to 10 g. To sweep a considerable pH range quite a lot of sample needs to be added: 20 g. This means the salinity of the solution in the titration vessel will change due to dilution with the titrant solution. For this reason, the salinity of the titrant solution needs to be given via the parameter `S_titrant`. However, we assume the titrant does not contain borate, sulfate or fluoride, that is why we do not set the flag `seawater_titrant` to `TRUE`.

```
> ae <- titration(ae_init, mass_sample = 0.01, mass_titrant = 0.02,  
+   conc_titrant = 0.01, S_titrant = 0.5, steps = 100)
```

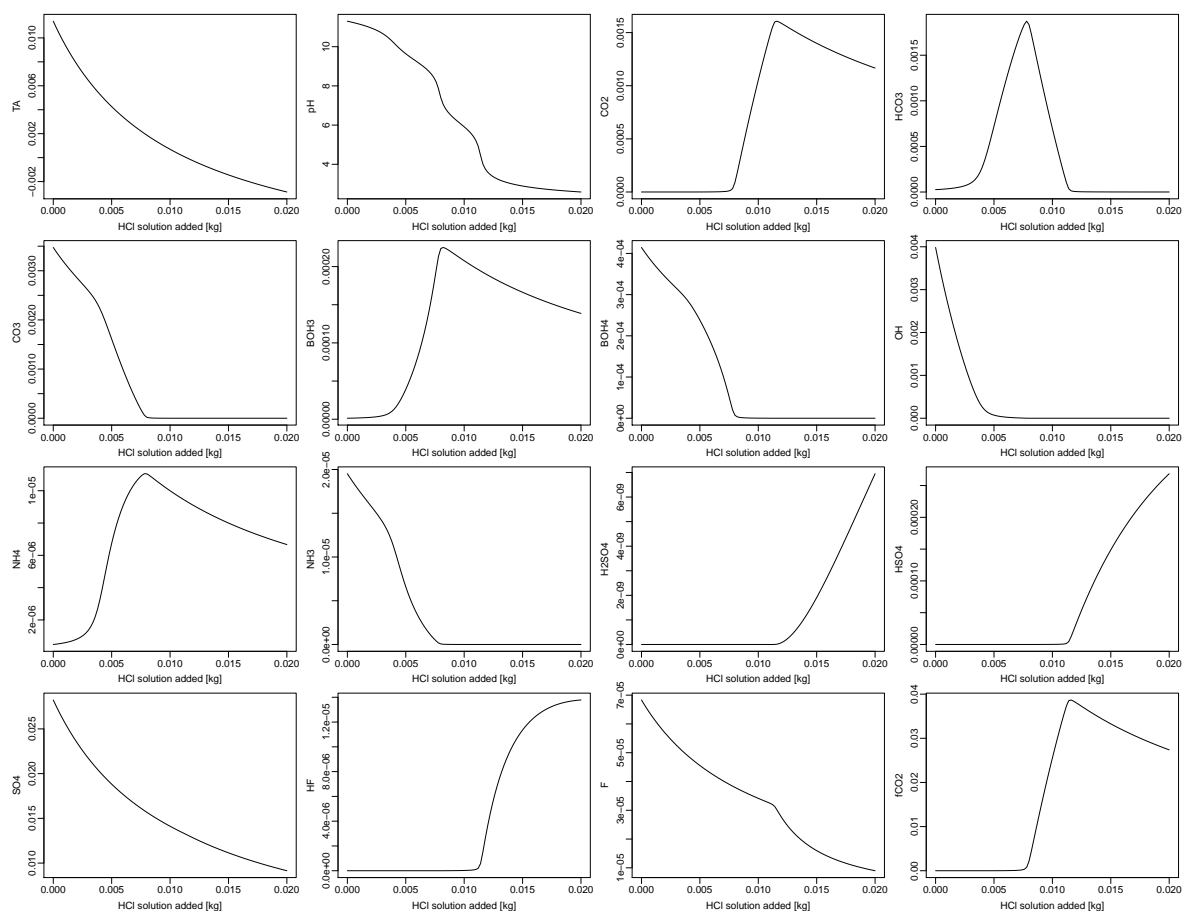
To get a quick overview, all elements of the obtained *aquaenv* object can be plotted

```
> plot(ae, xval = ae$delta_mass_titrant, xlab = "HCl solution added [kg]",  
+   mfrow = c(10, 10), newdevice = FALSE)
```



Then, a selection of elements can be plotted as a function of the added titrant mass,

```
> what <- c("TA", "pH", "CO2", "HCO3", "CO3", "BOH3", "BOH4", "OH",
+           "NH4", "NH3", "H2SO4", "HSO4", "SO4", "HF", "F", "fCO2")
> plot(ae, xval = ae$delta_mass_titrant, xlab = "HCl solution added [kg]",
+       what = what, size = c(12, 8), mfrow = c(4, 4), newdevice = FALSE)
```

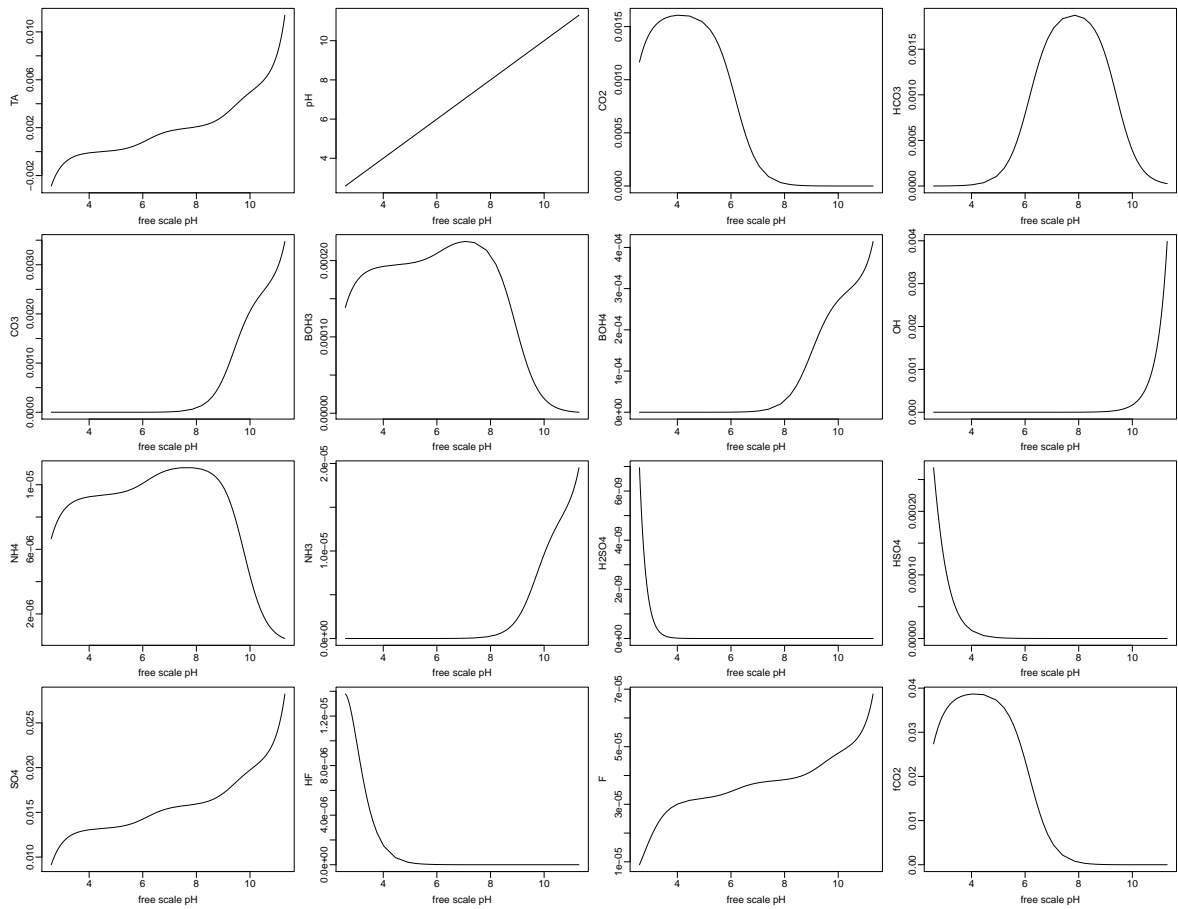


titrant concentration offset, or the moles of added titrant

```
> plot(ae, xval = ae$delta_conc_titrant, xlab = "[HCl] offset added [mol/kg-soln]",
+       what = what, size = c(14, 10), mfrow = c(4, 4))
> plot(ae, xval = ae$delta_moles_titrant, xlab = "HCl added [mol]",
+       what = what, size = c(14, 10), mfrow = c(4, 4))
```

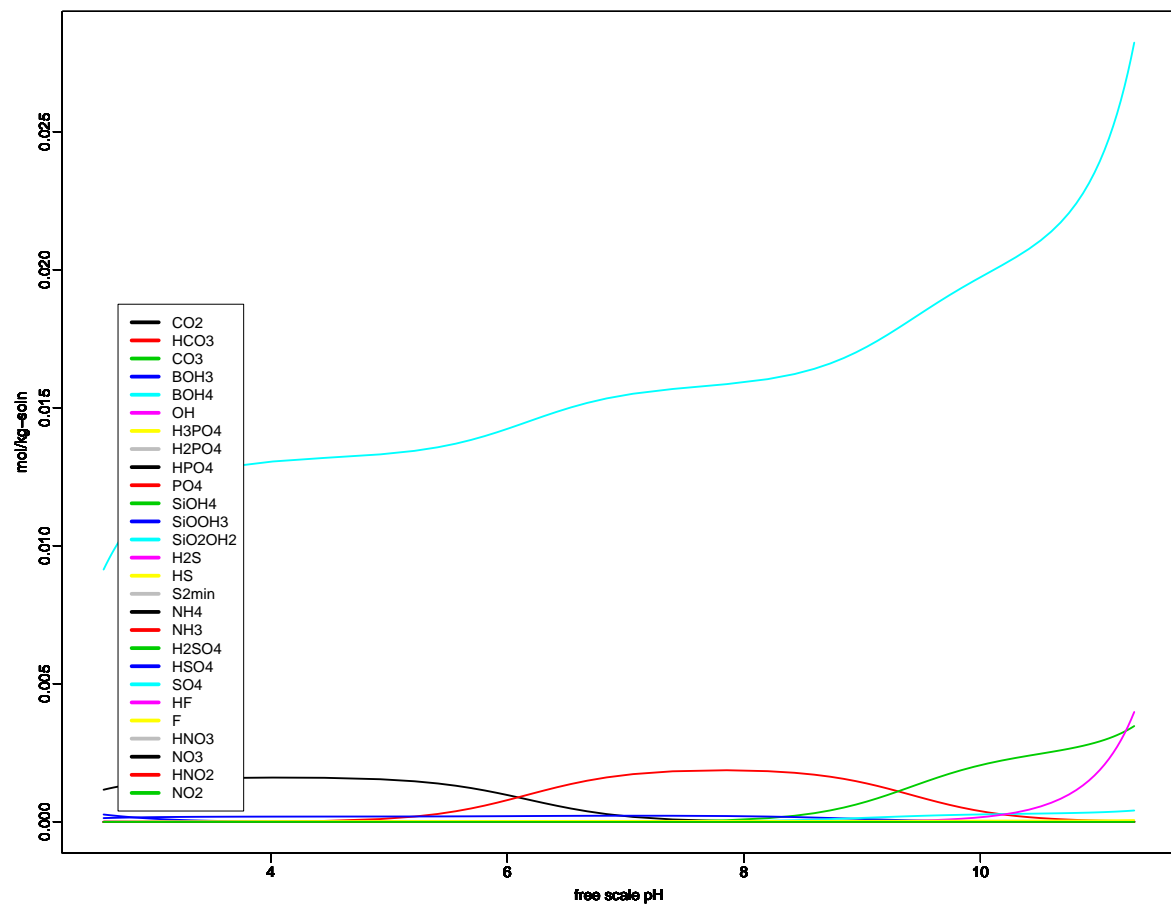
However, it is also possible to plot this selection of elements against other variables, e.g., the calculated free scale pH

```
> plot(ae, xval = ae$pH, xlab = "free scale pH", what = what, size = c(12,
+ 8), mfrow = c(4, 4), newdevice = FALSE)
```



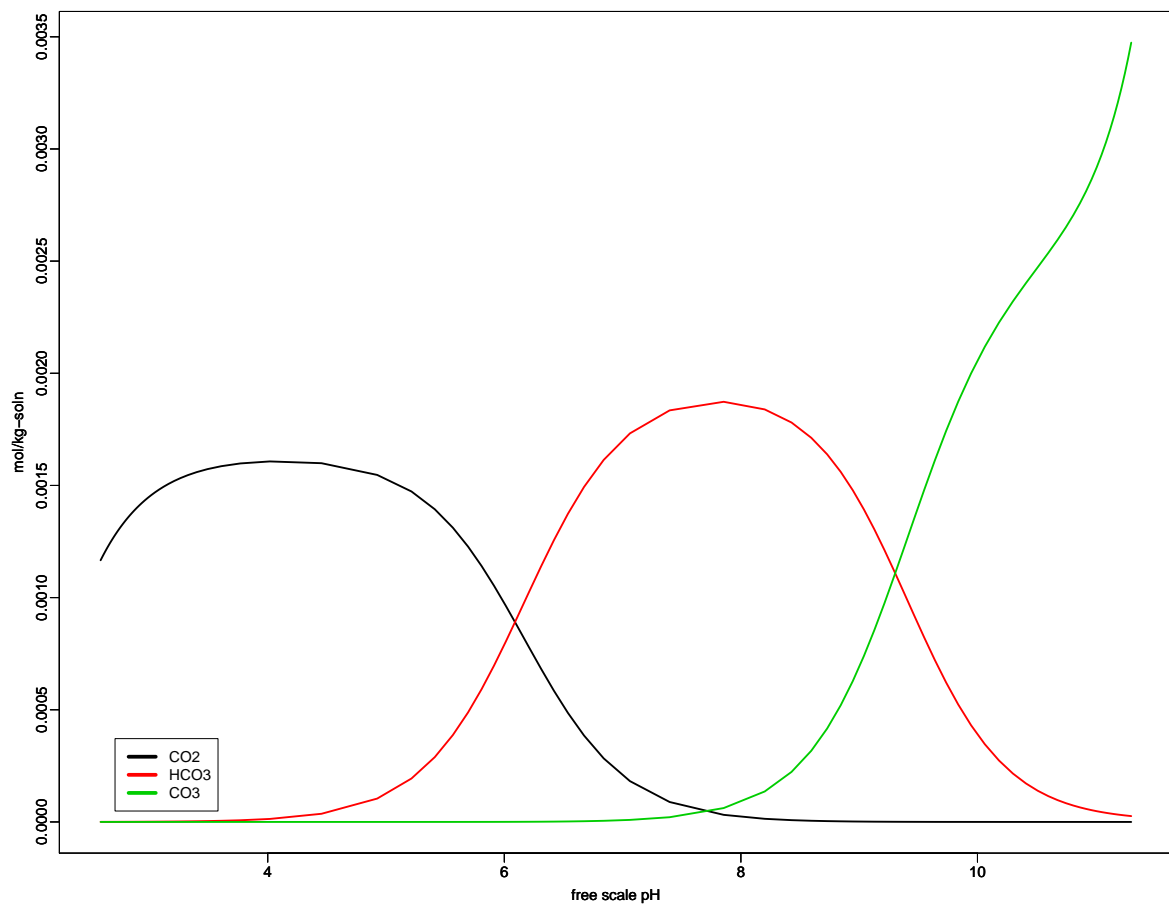
As mentioned earlier, the function `plot.aquaenv` offers the possibility of creating bjerrum plots from objects obtained with `titration`. The simplest way to do that is (remember the `newdevice=FALSE` is just needed to produce plots that are nicely woven into this vignette)

```
> plot(ae, bjerrum = TRUE, newdevice = FALSE)
```

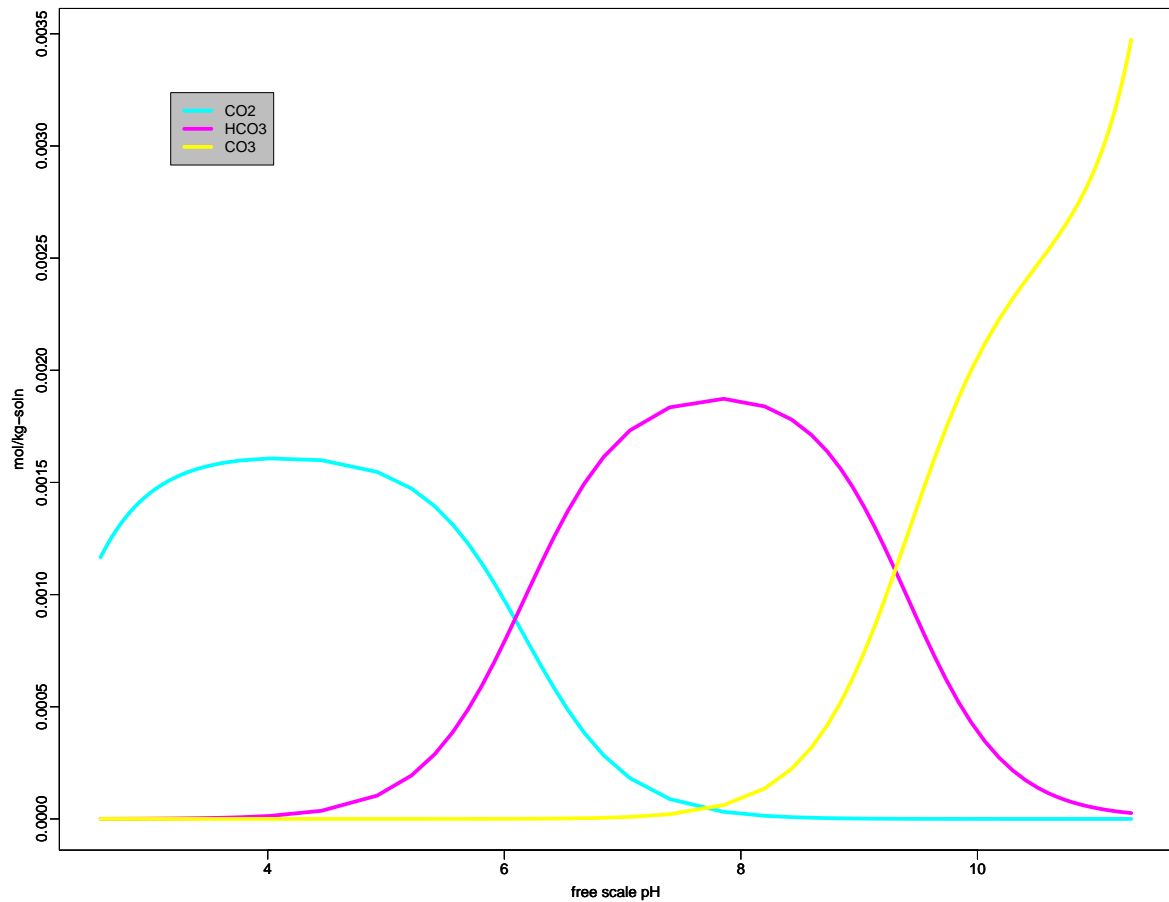
Or just select a few elements

```
> what <- c("CO2", "HCO3", "CO3")
> plot(ae, what = what, bjerrum = TRUE, newdevice = FALSE)
```



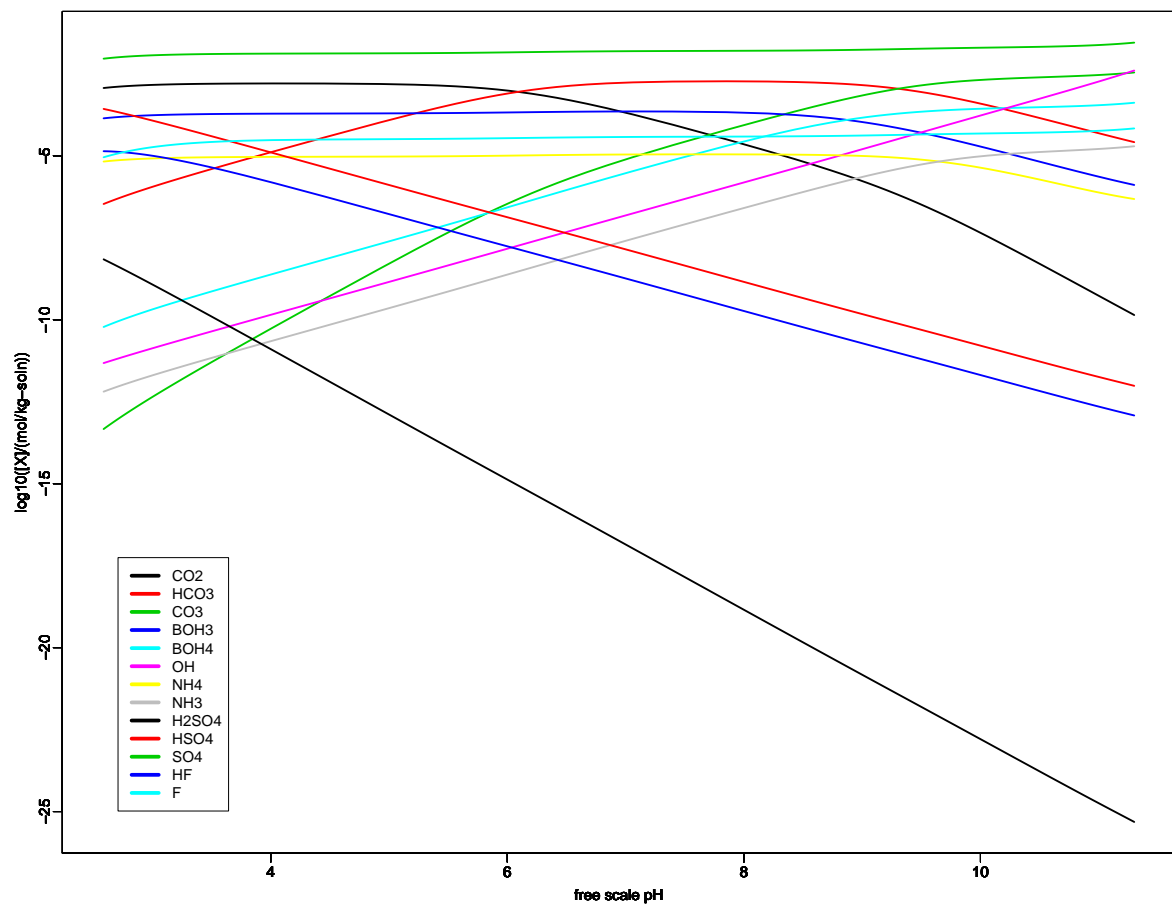
Again, the plotting style can be customized

```
> plot(ae, what = what, bjerrum = TRUE, lwd = 4, palette = c("cyan",
+   "magenta", "yellow"), bg = "gray", legendinset = 0.1, legendposition = "topleft",
+   newdevice = FALSE)
```



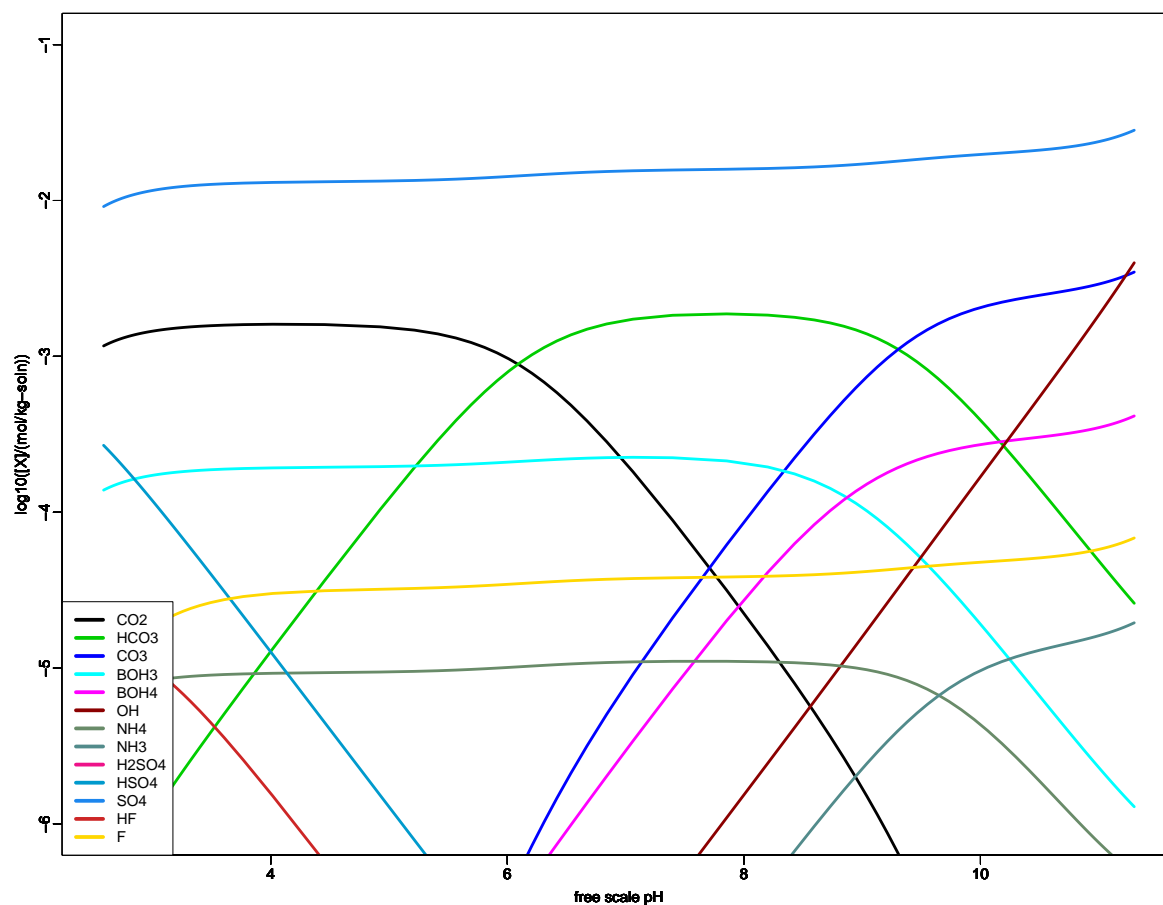
However, generally Bjerrum plots are done on the log scale. This can be accomplished using the flag `log`

```
> what <- c("CO2", "HCO3", "CO3", "BOH3", "BOH4", "OH", "NH4",
+           "NH3", "H2SO4", "HSO4", "SO4", "HF", "F")
> plot(ae, what = what, bjerrum = TRUE, log = TRUE, newdevice = FALSE)
```



Furthermore, we can zoom in to the region of most interest to marine scientists

```
> plot(ae, what = what, bjerrum = TRUE, log = TRUE, ylim = c(-6,
+   -1), legendinset = 0, lwd = 3, palette = c(1, 3, 4, 5, 6,
+   colors()[seq(100, 250, 6)]), newdevice = FALSE)
```



3.4.2 Titration with NaOH

Similar to the titration with HCl, also a titration with NaOH can be simulated (Please note that due to runtime constraints of the vignette for this package, this titration simulation is not run during construction of the vignette and, therefore, no resulting plots of the following code chunks is included. The user can generate the plots by extracting and executing the relevant code-chunks.)

```
> ae <- titration(aquaenv(S = 35, t = 15, SumCO2 = 0.0035, SumNH4 = 2e-05,
+   pH = 2), mass_sample = 0.01, mass_titrant = 0.02, conc_titrant = 0.01,
+   S_titrant = 0.5, steps = 50, type = "NaOH")
```

Plotting everything

```
> plot(ae, xval = ae$delta_mass_titrant, xlab = "NaOH solution added [kg]",
+   mfrow = c(10, 10))
```

Plotting selectively

```
> what <- c("TA", "pH", "CO2", "HCO3", "CO3", "BOH3", "BOH4", "OH",
+   "NH4", "NH3", "H2SO4", "HSO4", "SO4", "HF", "F", "fCO2")
```

```
> plot(ae, xval = ae$delta_mass_titrant, xlab = "NaOH solution added [kg]",
+       what = what, size = c(12, 8), mfrow = c(4, 4))
> plot(ae, xval = ae$pH, xlab = "free scale pH", what = what, size = c(12,
+       8), mfrow = c(4, 4))
```

Bjerrum plots

```
> what <- c("CO2", "HCO3", "CO3")
> plot(ae, what = what, bjerrum = TRUE, newdevice = FALSE)

> what <- c("CO2", "HCO3", "CO3", "BOH3", "BOH4", "OH", "NH4",
+       "NH3", "H2SO4", "HSO4", "SO4", "HF", "F")
> plot(ae, what = what, bjerrum = TRUE, log = TRUE, ylim = c(-6,
+       -1), legendinset = 0, lwd = 3, palette = c(1, 3, 4, 5, 6,
+       colors()[seq(100, 250, 6)]), newdevice = FALSE)
```

3.4.3 Titration with a titrant with high concentrations and a large sample volume - classical Bjerrum plots

The Bjerrum plots created in the previous two sections do not really look like the classical textbook ones. This is because we simulated a titration with a small sample volume and a titrant with low concentrations. As a result the total concentrations like, e.g., total carbonate decreased due to dilution. In simulating a titration with a rather large volume and a titrant with high concentrations the volume and salinity corrections do not matter any more and graphs known from textbooks (e.g. [Zeebe and Wolf-Gladrow 2001](#)) are produced.

```
> ae <- titration(aquaenv(S = 35, t = 15, SumCO2 = 0.0035, SumNH4 = 2e-05,
+       pH = 11.3), mass_sample = 100, mass_titrant = 0.5, conc_titrant = 3,
+       S_titrant = 0.5, steps = 100)
```

Plotting everything

```
> plot(ae, xval = ae$delta_mass_titrant, xlab = "HCl solution added [kg]",
+       mfrow = c(10, 10))
```

Plotting selectively and with different elements for xval

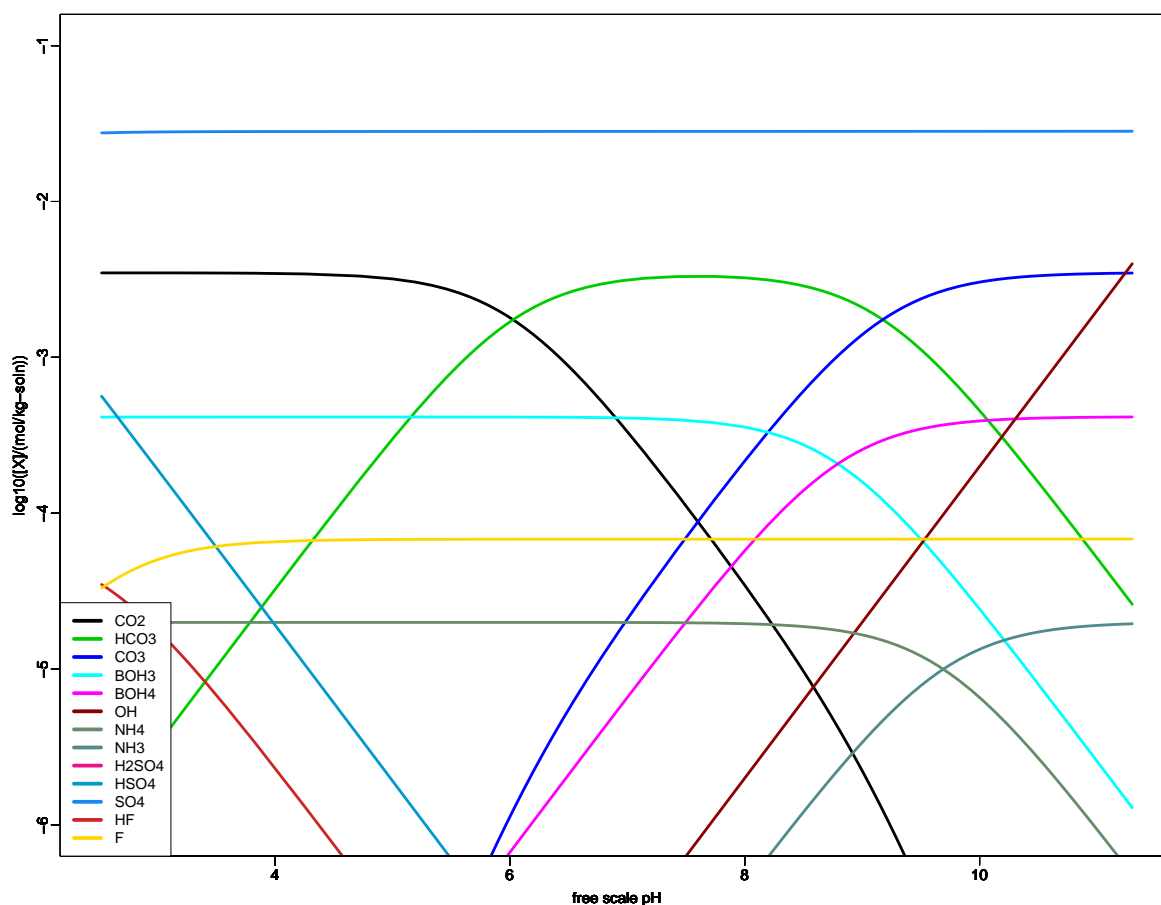
```
> what <- c("TA", "pH", "CO2", "HCO3", "CO3", "BOH3", "BOH4", "OH",
+       "NH4", "NH3", "H2SO4", "HSO4", "SO4", "HF", "F", "fCO2")
> plot(ae, xval = ae$delta_mass_titrant, xlab = "HCl solution added [kg]",
+       what = what, size = c(12, 8), mfrow = c(4, 4))
> plot(ae, xval = ae$pH, xlab = "free scale pH", what = what, size = c(12,
+       8), mfrow = c(4, 4))
> plot(ae, xval = ae$delta_conc_titrant, xlab = "[HCl] offset added [mol/kg-soln]",
+       what = what, size = c(12, 8), mfrow = c(4, 4))
> plot(ae, xval = ae$delta_moles_titrant, xlab = "HCl added [mol]",
+       what = what, size = c(12, 8), mfrow = c(4, 4))
```

Creating different kinds of Bjerrum plots

```
> plot(ae, bjerrum = TRUE)
> what <- c("CO2", "HCO3", "CO3")
> plot(ae, what = what, bjerrum = TRUE)
> plot(ae, what = what, bjerrum = TRUE, lwd = 4, palette = c("cyan",
+   "magenta", "yellow"), bg = "gray", legendinset = 0.1, legendposition = "topleft")
> what <- c("CO2", "HCO3", "CO3", "BOH3", "BOH4", "OH", "NH4",
+   "NH3", "H2SO4", "HSO4", "SO4", "HF", "F")
> plot(ae, what = what, bjerrum = TRUE, log = TRUE)
```

and the classical textbook one

```
> what <- c("CO2", "HCO3", "CO3", "BOH3", "BOH4", "OH", "NH4",
+   "NH3", "H2SO4", "HSO4", "SO4", "HF", "F")
> plot(ae, what = what, bjerrum = TRUE, log = TRUE, ylim = c(-6,
+   -1), legendinset = 0, lwd = 3, palette = c(1, 3, 4, 5, 6,
+   colors()[seq(100, 250, 6)]), newdevice = FALSE)
```



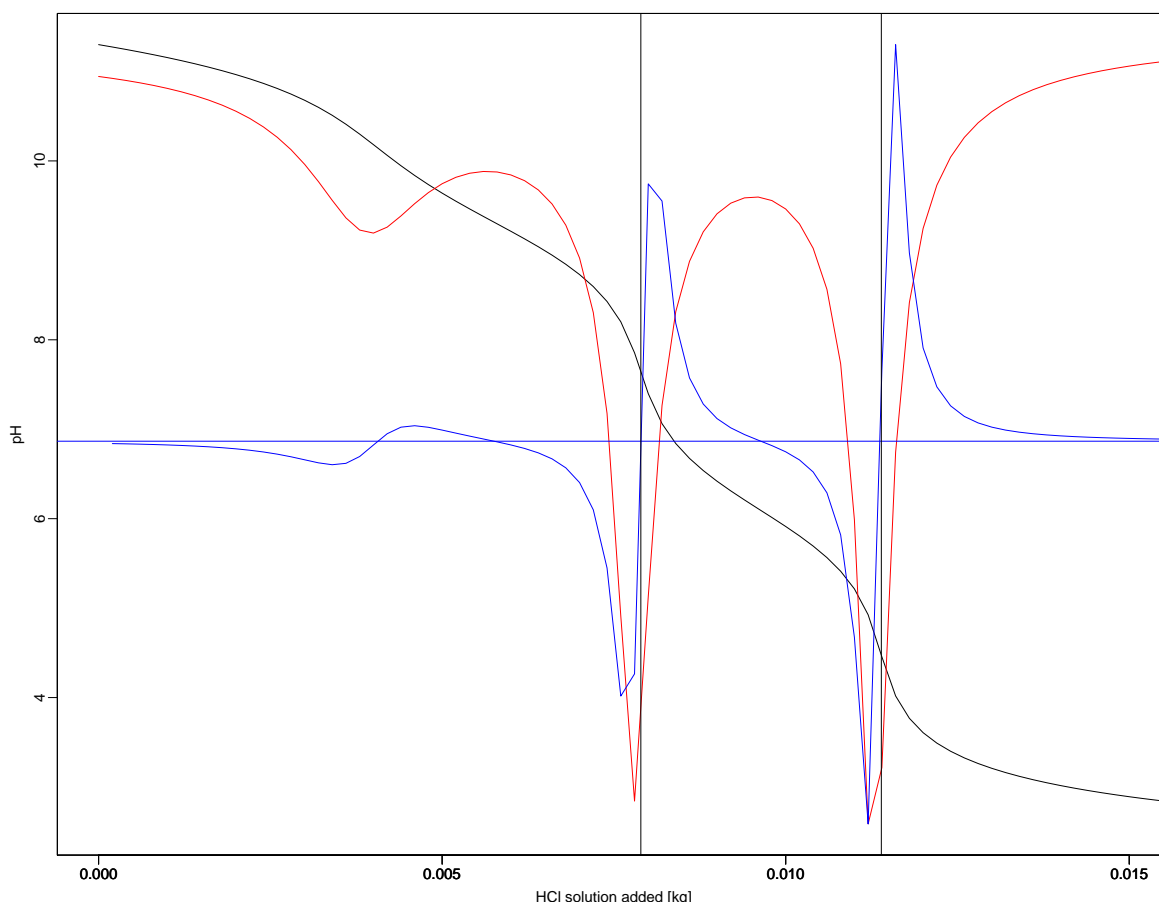
3.5 Calculating information from titration curves: the function `TAfit`

3.5.1 A little theory

While titrating a sample of natural seawater with HCl there one sees two clear equivalence points ([Dickson 1981](#)) The second equivalence point is the equivalence point of total alkalinity and the difference between the second and the first equivalence point signifies the total amount of $\sum \text{CO}_2$ of the sample [Hansson and Jagner \(1973\)](#).

This can be illustrated with **AquaEnv**. The respective titration curve can be plotted, together with its first and second derivative. Furthermore, the equivalence points can be marked with vertical lines (Please note that for a titrant concentration of 0.01 mol/kg-solution and 0.01 kg of sample, the value of the concentration (in mol/kg-solution) of total alkalinity and total carbonate equals the value of the total amount (in mol)).

```
> ae_init <- aquaenv(S = 35, t = 15, SumCO2 = 0.0035, SumNH4 = 2e-05,
+   pH = 11.3)
> ae <- titration(ae_init, mass_sample = 0.01, mass_titrant = 0.02,
+   conc_titrant = 0.01, S_titrant = 0.5, steps = 100)
> plot(ae, xval = ae$delta_mass_titrant, xlab = "HCl solution added [kg]",
+   what = "pH", xlim = c(0, 0.015), newdevice = FALSE)
> par(new = TRUE)
> plot(ae$delta_mass_titrant[1:100], diff(ae$pH), type = "l", col = "red",
+   xlim = c(0, 0.015), ylab = "", xlab = "", yaxt = "n")
> par(new = TRUE)
> plot(ae$delta_mass_titrant[2:100], diff(diff(ae$pH)), type = "l",
+   col = "blue", xlim = c(0, 0.015), ylab = "", xlab = "", yaxt = "n")
> abline(h = 0, col = "blue")
> abline(v = ae$TA[[1]])
> abline(v = ae$TA[[1]] - ae$SumCO2[[1]])
```

Following classical chemical textbooks (e.g. [Skoog and West 1982](#)), one can determine $[TA]$ and $[\sum CO_2]$ of a sample by graphically determining those equivalence points. However, there is no mechanistic understanding of the contents of the solution involved in doing so.

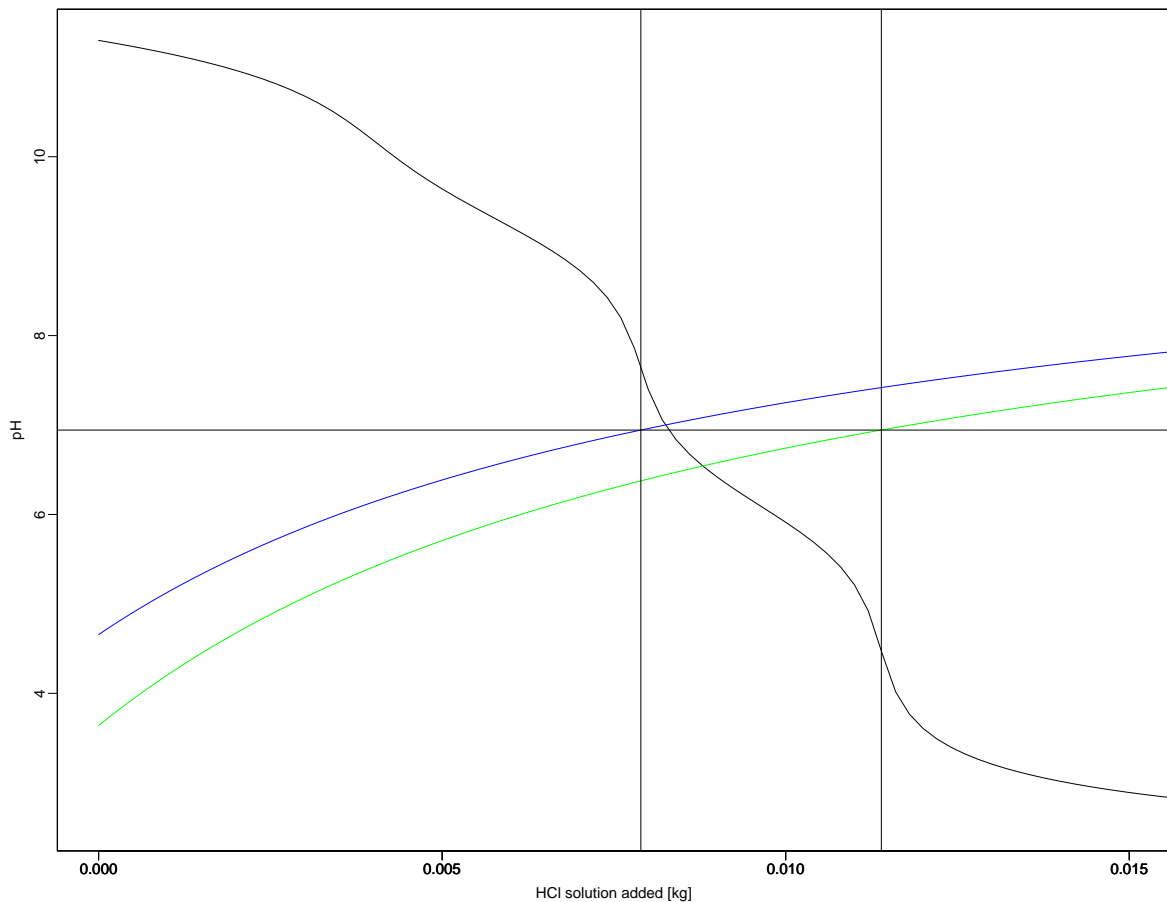
Other methods, called “Gran evaluations” ([Gran 1952](#); [Hansson and Jagner 1973](#); [Dickson 1981](#); [Haraldsson, Anderson, Hasselov, Hulth, and Olsson 1997](#); [Anderson, Turner, Wedborg, and Dyrssen 1999](#)), try to linearize the mechanistic model of what is going on in the solution during titration. They define the so called linear “Gran functions” and try to find their roots to determine the equivalence points. We will illustrate that by plotting the Gran functions F_0 (blue) and F_2 ($-F_1$, green) and again mark the equivalence points with vertical lines. The $y=0$ line for the Gran functions is indicated by a horizontal line

```
> plot(ae, xval = ae$delta_mass_titrant, xlab = "HCl solution added [kg]",
+      what = "pH", xlim = c(0, 0.015), newdevice = FALSE)
> prot1 <- c()
> for (i in 1:length(ae$pH)) {
+   prot1 <- c(prot1, (10^-(ae$pH[[i]]) + ae$HSO4[[i]] + ae$HF[[i]] +
+     ae$CO2[[i]] - ae$CO3[[i]] - ae$BOH4[[i]] - ae$OH[[i]]))
+ }
> par(new = TRUE)
> plot(ae$delta_mass_titrant, prot1, type = "l", col = "blue",
```

```

+   xlim = c(0, 0.015), ylab = "", xlab = "", yaxt = "n", ylim = c(-0.015,
+   0.015))
> prot2 <- c()
> for (i in 1:length(ae$pH)) {
+   prot2 <- c(prot2, (10^-(ae$pH[[i]]) + ae$HSO4[[i]] + ae$HF[[i]] -
+   ae$HCO3[[i]] - 2 * ae$CO3[[i]] - ae$BOH4[[i]] - ae$OH[[i]]))
+ }
> par(new = TRUE)
> plot(ae$delta_mass_titrant, prot2, type = "l", col = "green",
+   xlim = c(0, 0.015), ylab = "", xlab = "", yaxt = "n", ylim = c(-0.015,
+   0.015))
> abline(v = ae$TA[[1]])
> abline(v = ae$TA[[1]] - ae$SumCO2[[1]])
> abline(h = 0)

```

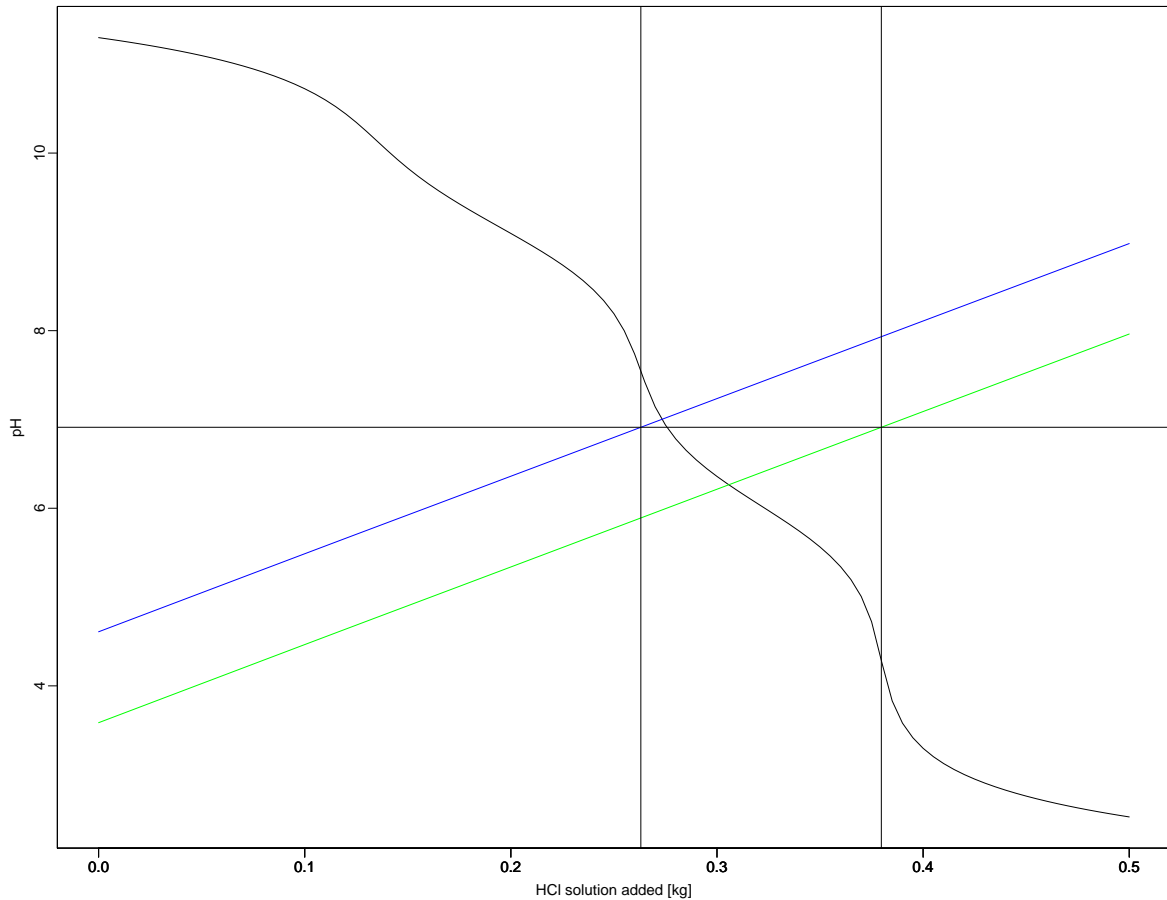


One can see that the Gran functions actually are not linear. This is due to volume and salinity change effects during the titration. This can be overcome by either employing “modified Gran functions” (see [Haraldsson *et al.* 1997](#)) that correct for the volume changes or by using a titration with a titrant with high concentrations and a large sample volume (Please note that here the value of the concentration of total alkalinity and total carbonate does not equal their total amount and need to be converted with the factor 100/3)

```

> ae <- titration(aquaenv(S = 35, t = 15, SumCO2 = 0.0035, SumNH4 = 2e-05,
+   pH = 11.3), mass_sample = 100, mass_titrant = 0.5, conc_titrant = 3,
+   S_titrant = 0.5, steps = 100)
> plot(ae, xval = ae$delta_mass_titrant, xlab = "HCl solution added [kg]",
+   what = "pH", xlim = c(0, 0.5), newdevice = FALSE)
> prot1 <- c()
> for (i in 1:length(ae$pH)) {
+   prot1 <- c(prot1, (10^-(ae$pH[[i]]) + ae$HSO4[[i]] + ae$HF[[i]] +
+     ae$CO2[[i]] - ae$CO3[[i]] - ae$BOH4[[i]] - ae$OH[[i]]))
+ }
> par(new = TRUE)
> plot(ae$delta_mass_titrant, prot1, type = "l", col = "blue",
+   xlim = c(0, 0.5), ylab = "", xlab = "", yaxt = "n", ylim = c(-0.015,
+     0.015))
> prot2 <- c()
> for (i in 1:length(ae$pH)) {
+   prot2 <- c(prot2, (10^-(ae$pH[[i]]) + ae$HSO4[[i]] + ae$HF[[i]] -
+     ae$HCO3[[i]] - 2 * ae$CO3[[i]] - ae$BOH4[[i]] - ae$OH[[i]]))
+ }
> par(new = TRUE)
> plot(ae$delta_mass_titrant, prot2, type = "l", col = "green",
+   xlim = c(0, 0.5), ylab = "", xlab = "", yaxt = "n", ylim = c(-0.015,
+     0.015))
> abline(v = (ae$TA[[1]] * 100/3))
> abline(v = ((ae$TA[[1]] - ae$SumCO2[[1]]) * 100/3))
> abline(h = 0)

```



Another proposed method of determining $[TA]$ and $[\sum CO_2]$ is to not only determine the two equivalence points, but to fit the whole titration curve with a theoretical titration curve based on a mechanistic model of what is going on in the solution during the titration (Dickson 1981; DOE 1994; Anderson *et al.* 1999). The function `titration` of **AquaEnv** provides exactly such a theoretical titration curve and the function `TAfit` makes use of this fact to determine $[TA]$ and $[\sum CO_2]$ of a sample by non linear curve fitting.

3.5.2 Determining $[TA]$ and $[\sum CO_2]$ by non linear curve fitting

3.5.2.1 Proof of concept

First, a proof of concept will show that the function `TAfit` is implemented consistently. Some "data" can be generated with the `titration` function.

```
> initial_ae <- aquaenv(S = 35, t = 15, SumCO2 = 0.002, TA = 0.0022)
> ae <- titration(initial_ae, mass_sample = 0.01, mass_titrant = 0.003,
+   conc_titrant = 0.01, S_titrant = 0.5, steps = 20)
```

Now, the input data for the `TAfit` routine can be generated: a table with the added mass of the titrant and the resulting free scale pH

```
> titcurve <- cbind(ae$delta_mass_titrant, ae$pH)
```

Note that For the `TAfit` all total quantities except `SumCO2` (`SumNH4`, `SumH2S`, `SumH3PO4`, `SumSiOH4`, `SumHNO3`, `SumHNO2`, `SumBOH3`, `SumH2SO4`, `SumHF`) need to be known. However, the latter three can be calculated from salinity as it is done in this example.

```
> fit1 <- TAfit(initial_ae, titcurve, conc_titrant = 0.01, mass_sample = 0.01,
+             S_titrant = 0.5)
> fit1
```

```
$TA
[1] 0.0022
attr(,"unit")
[1] "mol/kg-soln"
```

```
$SumCO2
[1] 0.002
attr(,"unit")
[1] "mol/kg-soln"
```

```
$sumofsquares
[1] 0
```

Thus, we see that `TAfit` calculates the correct `SumCO2` and `TA` values.

Trying the [Lueker, Dickson, and Keeling \(2000\)](#) (`K_CO2` and `K_HCO3`) and [Perez and Fraga \(1987b\)](#) (`K_HF`) values. (Please note that due to runtime constraints of the vignette for this package, this calculation is not run during construction of the vignette. The user can perform the calculation by extracting and executing the relevant code-chunks.)

```
> initial_ae_ <- aquaenv(S=35, t=15, SumCO2=0.002, TA=0.0022,
+                       k1k2="lueker", khf="perez")
> ae_ <- titration(initial_ae_, mass_sample=0.01, mass_titrant=0.003,
+                  conc_titrant=0.01,
+                  S_titrant=0.5, steps=20, k1k2="lueker", khf="perez")
> titcurve_ <- cbind(ae_$delta_mass_titrant, ae_$pH)
> fit1_ <- TAfit(initial_ae_, titcurve_, conc_titrant=0.01, mass_sample=0.01,
+               S_titrant=0.5, k1k2="lueker", khf="perez", verbose=TRUE)
> fit1_
```

`TAfit` can also take `E (V)` values as input variables, so we generate `E` values using `E0=0.4 V` and the Nernst equation. (But before that, we calculate a titration with a titrant with the same salinity as seawater such that `S` does not change during the titration. otherwise we would need to calculate the `S` profile for the titration extra to use it to convert to the total scale in the following step.) However, to do so we first need to convert our `pH` curve to the seawater `pH` scale. According to ([DOE 1994](#), p.7, ch.4, sop.3) and [Dickson, Sabine, and Christian \(2007\)](#), the Nernst equation relates `E` to the total proton concentration.

```
> ae <- titration(initial_ae, mass_sample = 0.01, mass_titrant = 0.003,
+   conc_titrant = 0.01, steps = 20, seawater_titrant = TRUE)
> titcurve <- cbind(ae$delta_mass_titrant, ae$pH)
> tottitcurve <- cbind(ae$pH, convert(ae$pH, "pHscale", "free2tot",
+   S = 35, t = 15))
> Etitcurve <- cbind(ae$delta_mass_titrant, (0.4 - ((PhysChemConst$R/10) *
+   initial_ae$T/PhysChemConst$F) * log(10^-tottitcurve[, 2])))
```

Again, T_Afit can be executed, this time also calculating E₀. Note that the flag `verbose=TRUE` causes T_Afit to show the progress of the fitting procedure in a plot window. (Please note that due to runtime constraints of the vignette for this package, this calculation is not run during construction of the vignette. The user can perform the calculation by extracting and executing the relevant code-chunks.)

```
> fit2 <- Tfit(initial_ae, Etitcurve, conc_titrant = 0.01, mass_sample = 0.01,
+   Evals = TRUE, verbose = TRUE, seawater_titrant = TRUE)
> fit2
```

Furthermore, T_Afit can fit K_{CO2} as well, however, one single value for the whole titration curve is fitted, i.e. there is no correction for K_{CO2} changes due to changing S due to mixing with the titrant

```
> fit3 <- Tfit(initial_ae, titcurve, conc_titrant = 0.01, mass_sample = 0.01,
+   S_titrant = 0.5, K_CO2fit = TRUE)
> fit3
```

```
$TA
```

```
[1] 0.002201253
```

```
attr("unit")
```

```
[1] "mol/kg-soln"
```

```
$SumCO2
```

```
[1] 0.002002897
```

```
attr("unit")
```

```
[1] "mol/kg-soln"
```

```
$K_CO2
```

```
[1] 9.169194e-07
```

```
attr("unit")
```

```
[1] "mol/kg-soln"
```

```
attr("pH scale")
```

```
[1] "free"
```

```
$sumofsquares
```

```
[1] 0.0001173601
```

```
> initial_ae$K_CO2
```

```
[1] 9.385853e-07
attr("unit")
[1] "mol/kg-soln"
attr("pH scale")
[1] "free"
```

One can see that the fitted value for K_CO2 is not the same as the value in the initial *aquaenv* object, which is the "correct" value. That is, because during data creation K_CO2 changed along the course of the titration due to changes in salinity. Assuming that the titrant has the same salinity as the sample (and is made up of natural seawater, i.e. containing SumBOH4, SumH2SO4 and SumHF as functions of S), then the "correct" K_CO2 should be fitted. This can be accomplished in *TAfit* by not giving the argument *S_titrant* (i.e. assuming the titrant has the same salinity as the sample) and setting the flag *seawater_titrant* to *TRUE*

```
> ae <- titration(initial_ae, mass_sample = 0.01, mass_titrant = 0.003,
+   conc_titrant = 0.01, steps = 20, seawater_titrant = TRUE)
> titcurve <- cbind(ae$delta_mass_titrant, ae$pH)
> fit4 <- TAfit(initial_ae, titcurve, conc_titrant = 0.01, mass_sample = 0.01,
+   K_CO2fit = TRUE, seawater_titrant = TRUE)
> fit4
```

```
$TA
[1] 0.0022
attr("unit")
[1] "mol/kg-soln"
```

```
$SumCO2
[1] 0.002
attr("unit")
[1] "mol/kg-soln"
```

```
$K_CO2
[1] 9.385853e-07
attr("unit")
[1] "mol/kg-soln"
attr("pH scale")
[1] "free"
```

```
$sumofsquares
[1] 7.928052e-29
```

Furthermore, TA, SumCO2, K_CO2 and E0 can be fitted at the same time. (Please note that due to runtime constraints of the vignette for this package, this calculation is not run during construction of the vignette. The user can perform the calculation by extracting and executing the relevant code-chunks.)

```
> Etitcurve <- cbind(titcurve[, 1], (0.4 - ((PhysChemConst$R/10) *
+   initial_ae$T/PhysChemConst$F) * log(10^-titcurve[, 2])))
```

```
> fit5 <- TAFit(initial_ae, Etitcurve, conc_titrant = 0.01, mass_sample = 0.01,
+   K_CO2fit = TRUE, seawater_titrant = TRUE, Evals = TRUE)
> fit5
```

Sometimes, the obtained titration curve is not equally spaced on the x axis. TAFit can deal with such curves if the flag `equalspaced` is set to `FALSE`. (Please note that due to runtime constraints of the vignette for this package, this calculation is not run during construction of the vignette. The user can perform the calculation by extracting and executing the relevant code-chunks.)

```
> neqsptitcurve <- rbind(titcurve[1:9, ], titcurve[11:20, ])
> fit6 <- TAFit(initial_ae, neqsptitcurve, conc_titrant = 0.01,
+   mass_sample = 0.01, seawater_titrant = TRUE, equalspaced = FALSE,
+   verbose = TRUE, debug = TRUE)
> fit6
```

Finally, some "noise" is added to the generated data

```
> noisetitcurve <- titcurve * rnorm(length(titcurve), mean = 1,
+   sd = 0.01)
> fit7 <- TAFit(initial_ae, noisetitcurve, conc_titrant = 0.01,
+   mass_sample = 0.01, seawater_titrant = TRUE, verbose = FALSE,
+   debug = TRUE)
> fit7
```

```
$TA
[1] 0.00222553
attr("unit")
[1] "mol/kg-soln"
```

```
$SumCO2
[1] 0.002021651
attr("unit")
[1] "mol/kg-soln"
```

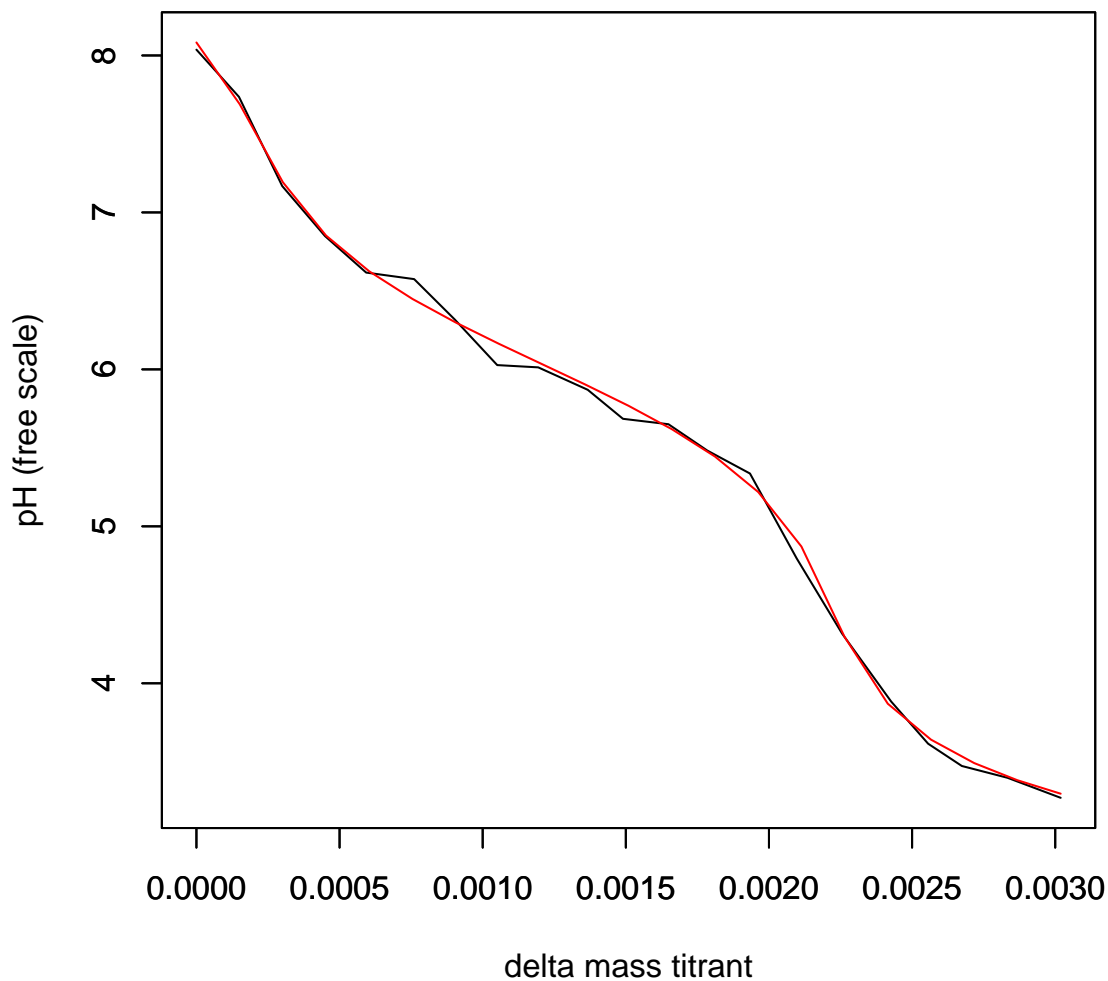
```
$sumofsquares
[1] 0.07271076
```

The flag `verbose=TRUE` (default is `FALSE`) prompts to show the traject of the fitting procedure in a plot window. However, each new fit is plotted over the first one and Sweave includes only the first plot in each code chunk in the resulting L^AT_EXfile. Therefore, we use the flag `debug=TRUE` to visualize the final fit

```
> ylim = range(noisetitcurve[, 2], calc)
> xlim = range(tit$delta_mass_titrant, noisetitcurve[, 1])
> plot(noisetitcurve[, 1], noisetitcurve[, 2], xlim = xlim, ylim = ylim,
+   type = "l", xlab = "delta mass titrant", ylab = "pH (free scale)")
```



```
> par(new = TRUE)
> plot(tit$delta_mass_titrant, calc, xlim = xlim, ylim = ylim,
+      type = "l", col = "red", xlab = "", ylab = "")
```



3.5.2.2 Test with generated data from literature

Dickson (1981) provided a synthetic dataset to test total alkalinity fitting programs. This dataset is included in **AquaEnv** as `sample_dickson1981`. Following quantities are given

```
> conc_titrant <- 0.3
> mass_sample <- 0.2
> S_titrant <- 14.835
```

```
> SumBOH3 <- 0.00042
> SumH2SO4 <- 0.02824
> SumHF <- 7e-05
```

Note that all concentrations are in mol/kg-solution and the mass of the sample is in kg. Note further that the salinity of the titrant has been calculated from its ionic strength of 0.3 mol/kg-soln.

In the original dataset as represented in `sample_dickson1981`, the mass of titrant is given in g which needs to be converted to kg

```
> sam <- cbind(sample_dickson1981[, 1]/1000, sample_dickson1981[,
+      2])
```

Then an attempt to recalculate the [TA] and $[\sum \text{CO}_2]$ values given in [Dickson \(1981\)](#) ([TA]=0.00245 mol/kg-soln and $[\sum \text{CO}_2]$ 0.00220 mol/kg-soln) can be done

```
> dicksonfit <- TAFit(aquaenv(S = 35, t = 25, SumBOH3 = SumBOH3,
+      SumH2SO4 = SumH2SO4, SumHF = SumHF), sam, conc_titrant, mass_sample,
+      S_titrant = S_titrant, debug = TRUE)
> dicksonfit
```

```
$TA
[1] 0.002464256
attr(,"unit")
[1] "mol/kg-soln"
```

```
$SumCO2
[1] 0.002191958
attr(,"unit")
[1] "mol/kg-soln"
```

```
$sumofsquares
[1] 0.01287255
```

This shows the fit is not accurate. Why is that so?

3.5.2.2.1 Does the salinity correction (`S_titrant`) matter?

Let us calculate a theoretical titration without salinity correction

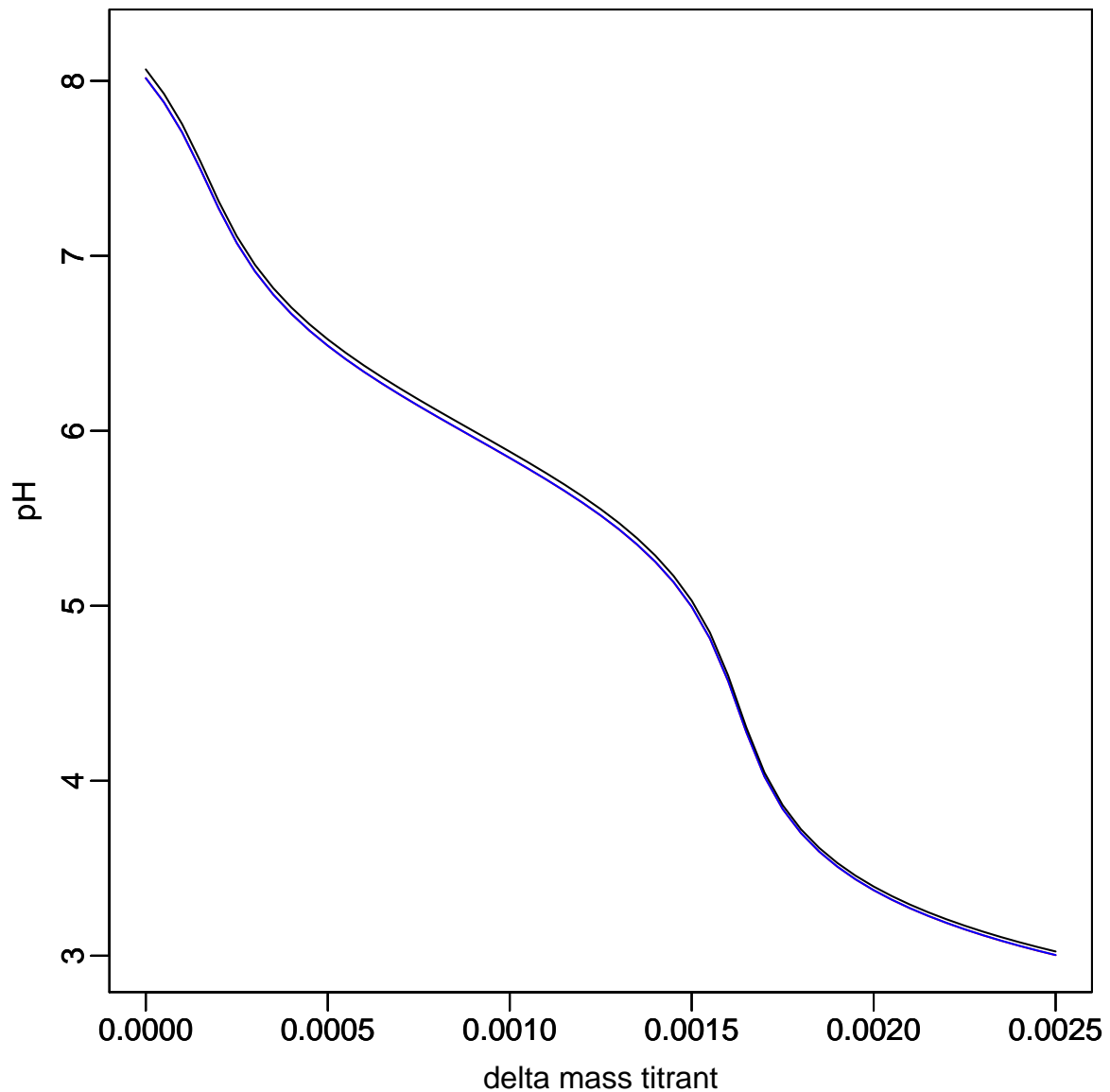
```
> dickson titration1 <- titration(aquaenv(S = 35, t = 25, SumCO2 = 0.0022,
+      SumBOH3 = SumBOH3, SumH2SO4 = SumH2SO4, SumHF = SumHF, TA = 0.00245),
+      mass_sample = mass_sample, mass_titrant = 0.0025, conc_titrant = conc_titrant,
+      steps = 50, type = "HCl")
```

and one with salinity correction

```
> dickson2titration2 <- titration(aquaenv(S = 35, t = 25, SumCO2 = 0.0022,  
+   SumBOH3 = SumBOH3, SumH2SO4 = SumH2SO4, SumHF = SumHF, TA = 0.00245),  
+   mass_sample = mass_sample, mass_titrant = 0.0025, conc_titrant = conc_titrant,  
+   S_titrant = S_titrant, steps = 50, type = "HCl")
```

Now the difference between both curves (in red and blue) and the “Dickson” curve (in black) can be visualized

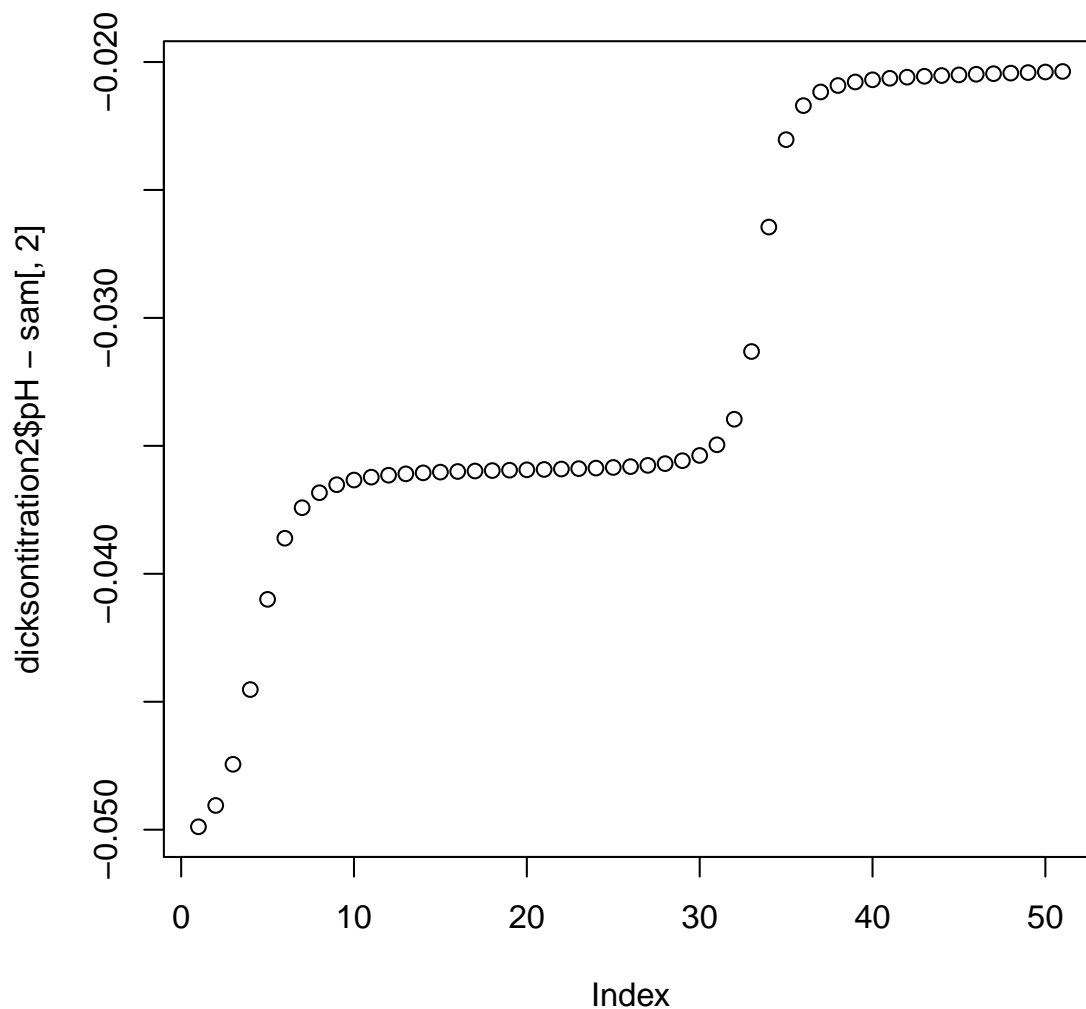
```
> plot(dickson2titration1, xval = dickson2titration1$delta_mass_titrant,  
+   what = "pH", xlim = c(0, 0.0025), ylim = c(3, 8.2), newdevice = FALSE,  
+   col = "red", xlab = "delta mass titrant")  
> par(new = TRUE)  
> plot(dickson2titration2, xval = dickson2titration2$delta_mass_titrant,  
+   what = "pH", xlim = c(0, 0.0025), ylim = c(3, 8.2), newdevice = FALSE,  
+   col = "blue", xlab = "")  
> par(new = TRUE)  
> plot(sam[, 1], sam[, 2], type = "l", xlim = c(0, 0.0025), ylim = c(3,  
+   8.2), xlab = "", ylab = "")
```



That means, the salinity correction makes no significant difference (the red and the blue curve cannot be discerned), because the relation between the total amount of sample and the added amount of titrant is very large: salinity only drops from 35 to 34.75105.

But there is an offset between the "Dickson" curve and our curve

```
> plot(dicksontitration2$pH - sam[, 2])
```



3.5.2.2.2 Does fitting K₂CO₃ as well improve the fit?

```
> dicksonfit2 <- TAfit(aquaenv(S = 35, t = 25, SumBOH3 = SumBOH3,
+   SumH2SO4 = SumH2SO4, SumHF = SumHF), sam, conc_titrant, mass_sample,
+   S_titrant = S_titrant, debug = TRUE, K_CO2fit = TRUE)
> dicksonfit2
```

```
$TA
[1] 0.002458081
attr(,"unit")
[1] "mol/kg-soln"
```

```
$SumCO2
[1] 0.002194006
attr("unit")
[1] "mol/kg-soln"
```

```
$K_CO2
[1] 1.030960e-06
attr("unit")
[1] "mol/kg-soln"
attr("pH scale")
[1] "free"
```

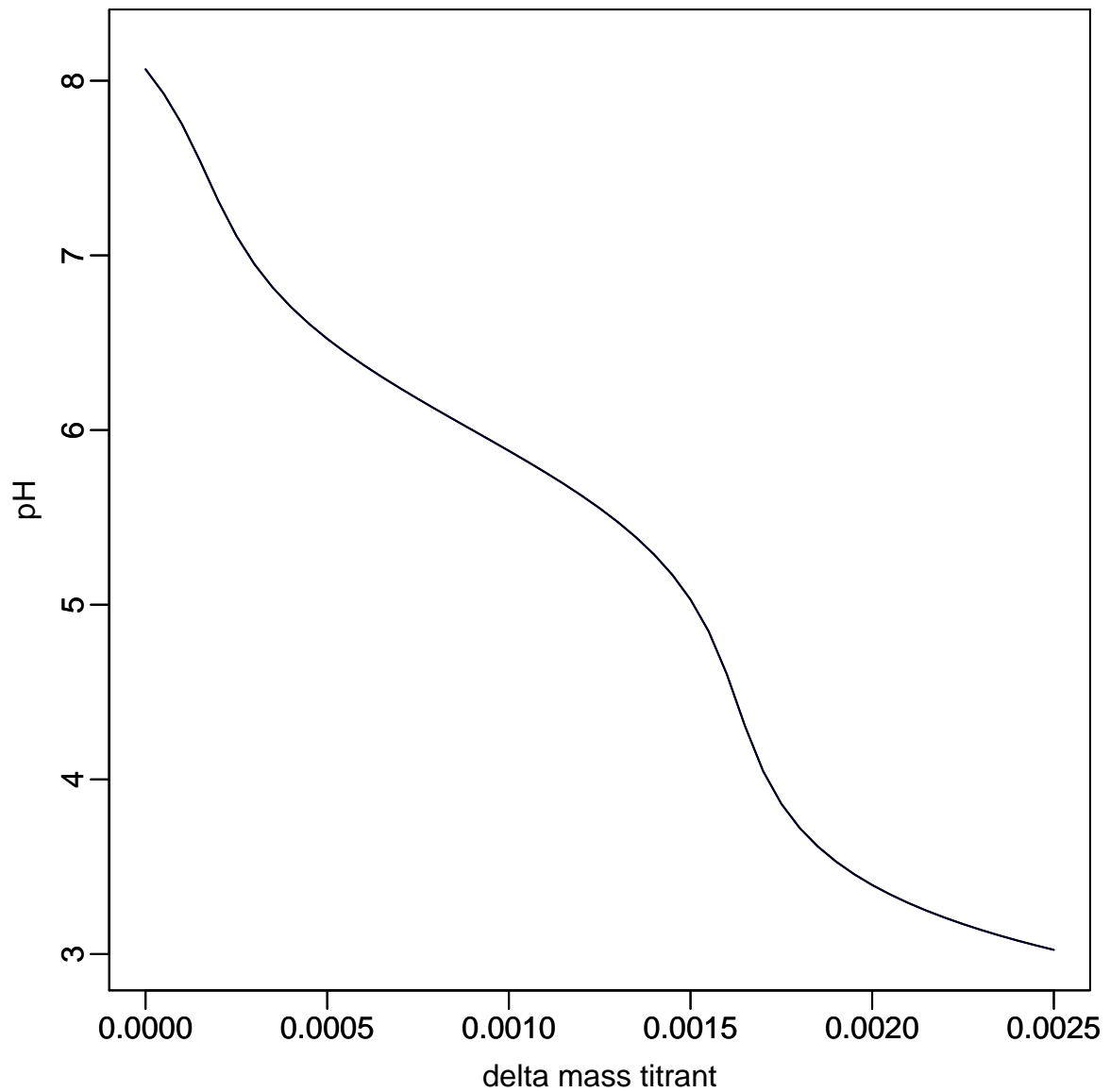
```
$sumofsquares
[1] 0.005724457
```

Yes it does, but it is not optimal yet.

There still remains one major difference between the calculations carried out in [Dickson \(1981\)](#) and the calculations in **AquaEnv**: [Dickson \(1981\)](#) uses fixed values for the equilibrium constants and does not calculate them as functions of temperature and salinity. Furthermore, the values that are used in [Dickson \(1981\)](#) are not exactly the same as are obtained in **AquaEnv** for the same salinity and temperature.

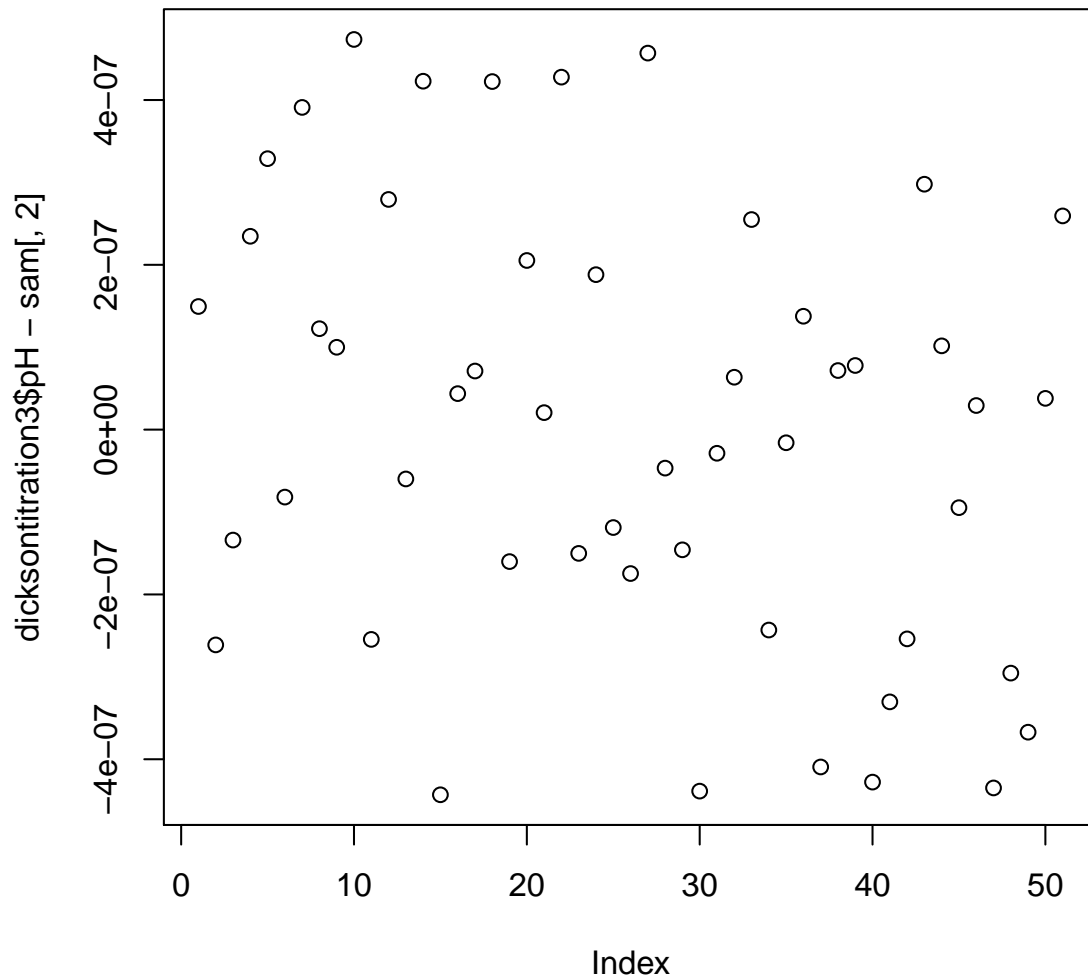
Let us calculate a theoretical titration curve employing exactly the same equilibrium constant values as used in [Dickson \(1981\)](#) and plot the result together with the “Dickson” curve

```
> dickson titration3 <- titration(aquaenv(S = 35, t = 25, SumCO2 = 0.0022,
+   SumBOH3 = SumBOH3, SumH2SO4 = SumH2SO4, SumHF = SumHF, TA = 0.00245,
+   k_w = 4.32e-14, k_co2 = 1e-06, k_hco3 = 8.2e-10, k_boh3 = 1.78e-09,
+   k_hso4 = (1/12.3), k_hf = (1/408)), mass_sample = mass_sample,
+   mass_titrant = 0.0025, conc_titrant = conc_titrant, steps = 50,
+   type = "HCl", S_titrant = S_titrant, k_w = 4.32e-14, k_co2 = 1e-06,
+   k_hco3 = 8.2e-10, k_boh3 = 1.78e-09, k_hso4 = (1/12.3), k_hf = (1/408))
> plot(dickson titration3, xval = dickson titration3$delta_mass_titrant,
+   what = "pH", xlim = c(0, 0.0025), ylim = c(3, 8.2), newdevice = FALSE,
+   col = "blue", xlab = "delta mass titrant")
> par(new = TRUE)
> plot(sam[, 1], sam[, 2], type = "l", xlim = c(0, 0.0025), ylim = c(3,
+   8.2), xlab = "", ylab = "")
```



Plotting the differences between both curves reveals that they are the same down to 1 $\mu\text{mol/kg}$ -soln.

```
> plot(dicksontitration3$pH - sam[, 2])
```



Calculating [TA] and $[\sum \text{CO}_2]$ using **TAfit** and exactly the same equilibrium constant values as used in [Dickson \(1981\)](#)

```
> dicksonfit3 <- TAfit(aquaenv(S = 35, t = 25, SumBOH3 = SumBOH3,
+   SumH2SO4 = SumH2SO4, SumHF = SumHF, k_w = 4.32e-14, k_co2 = 1e-06,
+   k_hco3 = 8.2e-10, k_boh3 = 1.78e-09, k_hso4 = (1/12.3), k_hf = (1/408)),
+   sam, conc_titrant, mass_sample, S_titrant = S_titrant, debug = TRUE,
+   k_w = 4.32e-14, k_co2 = 1e-06, k_hco3 = 8.2e-10, k_boh3 = 1.78e-09,
+   k_hso4 = (1/12.3), k_hf = (1/408))
> dicksonfit3

$TA
[1] 0.00245
attr(,"unit")
```



```
[1] "mol/kg-soln"
```

```
$SumCO2
```

```
[1] 0.0022
```

```
attr(,"unit")
```

```
[1] "mol/kg-soln"
```

```
$sumofsquares
```

```
[1] 3.279302e-12
```

reveals that now exactly the same values are calculated as are given in [Dickson \(1981\)](#).

4 Extending AquaEnv

It is very simple for the user to create own functions that use **AquaEnv** and extend its functionality. We will demonstrate that by creating simple analogons for the **AquaEnv** functions `titration` and `TAfit`.

The function `simpletitration` will take the following arguments

aquaenv an object of class *aquaenv*: minimal definition, contains all information about the system: S, t, p, total concentrations of nutrients etc

volume the volume of the (theoretical) titration vessel in l

amount the amount of titrant added in mol

steps the amount of steps the amount of titrant is added in

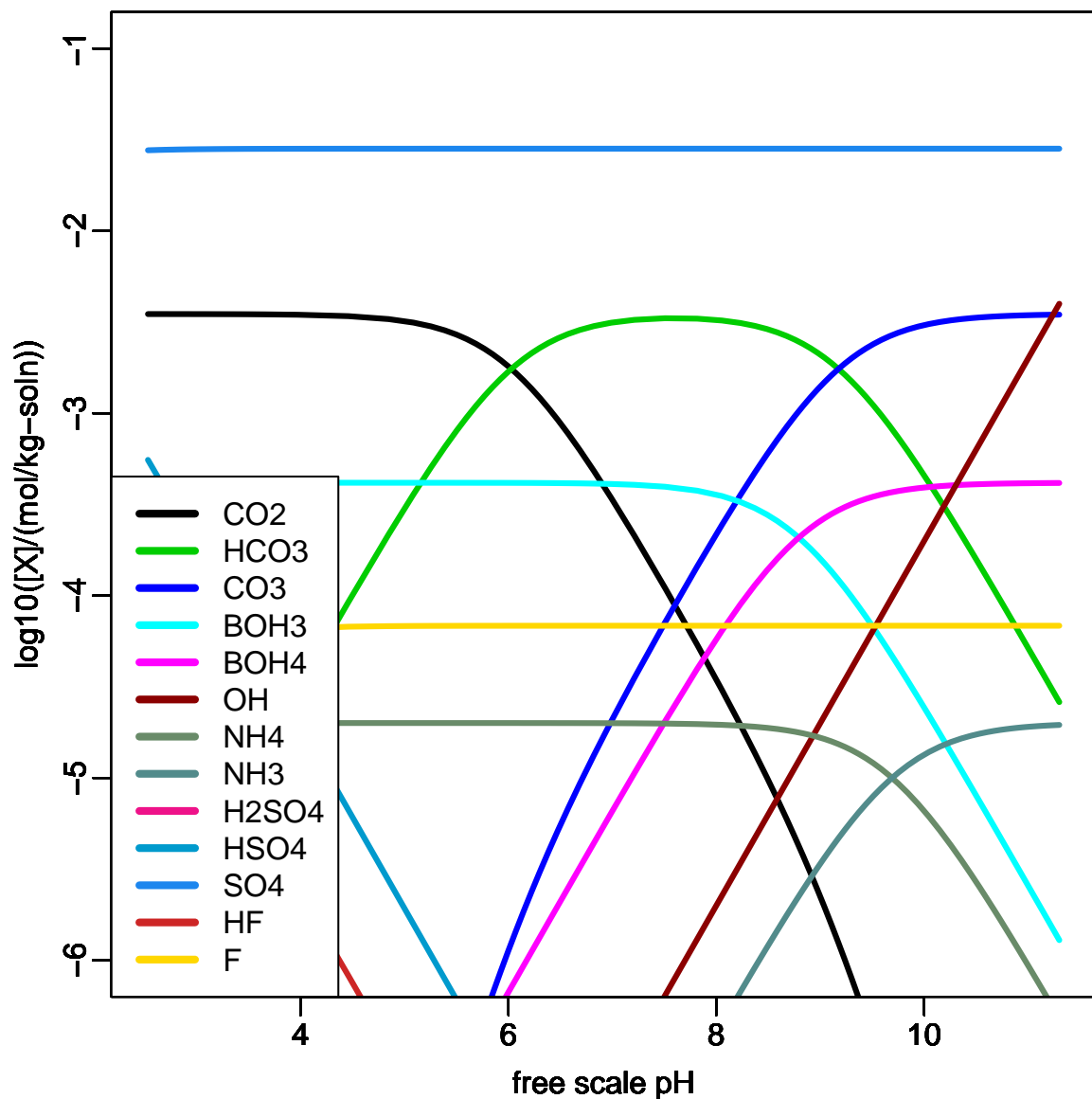
type the type of titrant: either "HCl" or "NaOH"

The function is defined as

```
> simpletitration <- function(aquaenv,                # an object of class aquaenv: minimal definition,
+                               # contains all information about the system:
+                               # T, S, d, total concentrations of nutrients etc
+                               volume,                # the volume of the (theoretical) titration vessel in l
+                               amount,                # the amount of titrant added in mol
+                               steps,                # the amount of steps the amount of titrant is added in
+                               type)                  # the type of titrant: either "HCl" or "NaOH"
+ {
+   directionTAchange <- switch(type, HCl = -1, NaOH = +1)
+   TAconcchangeperstep <- convert(((amount/steps)/volume), "conc", "molar2molin", aquaenv$t, aquaenv$S)
+
+   aquaenvtemp <- aquaenv
+
+   for (i in 1:steps)
+   {
+     TA <- aquaenvtemp$TA + (directionTAchange * TAconcchangeperstep)
+     aquaenvtemp <- aquaenv(ae=aquaenvtemp, TA=TA)
+     aquaenv <- merge(aquaenv, aquaenvtemp)
+   }
+
+   aquaenv[["DeltaCTitrant"]] <- convert((amount/volume)/steps*(1:(steps+1)),
+                                         "conc", "molar2molin", aquaenv$t, aquaenv$S)
+   return(aquaenv) # object of class aquaenv which contains a titration simulation
+ }
```

and can be used to create a bjerrum plot

```
> ae <- simpletitration(aquaenv(S = 35, t = 15, SumCO2 = 0.0035,
+   SumNH4 = 2e-05, pH = 11.3), volume = 100, amount = 1.5, steps = 100,
+   type = "HCl")
> what <- c("CO2", "HCO3", "CO3", "BOH3", "BOH4", "OH", "NH4",
+   "NH3", "H2SO4", "HSO4", "SO4", "HF", "F")
> plot(ae, what = what, bjerrum = TRUE, log = TRUE, ylim = c(-6,
+   -1), legendinset = 0, lwd = 3, palette = c(1, 3, 4, 5, 6,
+   colors()[seq(100, 250, 6)]), newdevice = FALSE)
```



The function `simpletitration` in turn can be used to create a simple analogon to `TAfit` with the arguments

ae an object of class *aquaenv*: minimal definition, contains all information about the system: S, t, p, total concentrations of nutrients etc

pHmeasurements a table containing the titration curve: basically a series of pH values (pH on free proton scale)

volume the volume of the titration vessel

amount the total amount of the titrant added

TAguess=0.0025 a first guess for [TA] and [SumCO₂] to be used as initial values for the optimization procedure

type="HCl" the type of titrant: either "HCl" or "NaOH"

defined as

```
> simpleTAfit <- function(ae,                                # an object of class aquaenv: minimal definition,
+                          # contains all information about the system:
+                          # T, S, d, total concentrations of nutrients etc
+                          pHmeasurements,                  # a table containing the titration curve:
+                          # basically a series of pH values (pH on free proton scale)
+                          volume,                          # the volume of the titration vessel
+                          amount,                          # the total amount of the titrant added
+                          TAguess=0.0025,                 # a first guess for [TA] and [SumCO2] to be used as
+                          # initial values for the optimization procedure
+                          type="HCl")                      # the type of titrant: either "HCl" or "NaOH"
+ {
+   ae$Na <- NULL # make sure ae gets cloned as "skeleton": cloneaquaenv determines "skeleton"
+               # TRUE or FALSE from the presence of a value for Na
+   residuals <- function(state)
+   {
+     ae$SumCO2 <- state[[1]]
+     pHcalc    <- simpletitration(aquaenv(ae=ae, TA=state[[2]]), volume=volume,
+                                   amount=amount, steps=(length(pHmeasurements)-1), type=type)$pH
+     residuals <- pHmeasurements-pHcalc
+     return(residuals)
+   }
+   require(minpack.lm)
+   out <- nls.lm(fn=residuals, par=c(TAguess, TAguess)) #guess for TA is also used as guess for SumCO2
+   result <- list(out$par[[2]], out$par[[1]], out$deviance)
+   attr(result[[1]], "unit") <- "mol/kg-soln"
+   attr(result[[2]], "unit") <- "mol/kg-soln"
+   names(result) <- c("TA", "SumCO2", "sumofsquares")
+   return(result) # a list of three values
+                 # ([TA] in mol/kg-solution, [SumCO2] in mol/kg-solution, sum of the squared residuals)
+ }
```

The function `simpleTAfit` can be used to calculate TA and SumCO₂

```
> pHmeasurements <- ae$pH
> fit <- simpleTAfit(aquaenv(S = 35, t = 15, SumNH4 = 2e-05), pHmeasurements,
+   volume = 100, amount = 1.5)
> fit
```

```
$TA
[1] 0.01139192
```

```
attr("unit")  
[1] "mol/kg-soln"
```

```
$SumC02  
[1] 0.0035  
attr("unit")  
[1] "mol/kg-soln"
```

```
$sumofsquares  
[1] 1.587199e-20
```

A Abbreviations for references used throughout the code and in the helpfiles

Atkins1996	Atkins (1996)
Boudreau1996	Boudreau (1996)
DOE1994	DOE (1994)
Dickson1979a	Dickson and Riley (1979a)
Dickson1981	Dickson (1981)
Dickson1984	Dickson (1984)
Dickson1987	Dickson and Millero (1987)
Dickson1990	Dickson (1990a)
Dickson2007	Dickson <i>et al.</i> (2007)
Emerson2008	Emerson and Hedges (2008)
Feistel2008	Feistel (2008)
Fofonoff1983	Fofonoff and Millard (1983)
Follows2006	Follows, Ito, and Dutkiewicz (2006)
Hofmann2008	Hofmann <i>et al.</i> (2008a)
Khoo1977	Khoo, Ramette, Culberson, and Bates (1977)
Lewis1998	Lewis and Wallace (1998)
Lueker2000	Lueker <i>et al.</i> (2000)
Millero1981	Millero and Poisson (1981)
Millero1988	Millero <i>et al.</i> (1988)
Millero1995	Millero (1995)
Millero1995a	Millero <i>et al.</i> (1995)
Millero2006	Millero, Graham, Huang, Bustos-Serrano, and Pierrot (2006)
Mucci1983	Mucci (1983)
Perez1987a	Perez and Fraga (1987a)
Riordan2005	Riordan <i>et al.</i> (2005)
Roy1993b	Roy <i>et al.</i> (1993b)
Sundquist1979	Sundquist, Plummer, and Wigley (1979)
Weiss1970	Weiss (1970)
Weiss1974	Weiss (1974)
Wischmeyer2003	Wischmeyer <i>et al.</i> (2003)
Zeebe2001	Zeebe and Wolf-Gladrow (2001)

B References for the elements of an object of class *aquaenv*

element	references
p, P, Pa, p	The relation between pressure and depth given in Fofonoff and Millard (1983) is used. The standard value for atmospheric pressure Pa at sea level as well as the definition of total pressure and gauge pressure is taken from Feistel (2008) .
Cl	DOE (1994, chapter 5, p. 11) , and Zeebe and Wolf-Gladrow (2001, p. 100, footnote 3)
I	DOE (1994, chapter 5, p. 13, 15) , Zeebe and Wolf-Gladrow (2001, p.12) , and Roy <i>et al.</i> (1993b, p.257) . Note that the approximation $I/(\text{mol/kg-solution}) \approx 0.0199201$ S is given in Millero (1982, p. 428.) . This relationship converted into mol/kg-H ₂ O and the last digits adjusted (from 0.0199201 to 0.019924) results in the formula used here.
density	Millero and Poisson (1981) and DOE (1994, chapter 5, p. 6f) .
Br, ClConc, Na, Mg, Ca, K, Sr	DOE (1994, chapter 5, p.11)
molal2molin	Roy <i>et al.</i> (1993b, p.257) , and DOE (1994, chapter 5, p. 15)
free2tot, tot2free	Dickson (1984, p.2302) , DOE (1994, chapter 4, p.16) , Zeebe and Wolf-Gladrow (2001, p.57, 261)
free2sws, tot2sws, sws2free, sws2tot	Dickson (1984, p.2303) , Zeebe and Wolf-Gladrow (2001, p.57)
K0_CO2	Weiss (1974) , DOE (1994, chapter 5, p. 13) (here it is stated that the unit is mol/(kg-solution*atm)), Millero (1995, p.663) , Zeebe and Wolf-Gladrow (2001, p.257)
K0_O2	derived from a formula for the oxygen saturation concentration in ml-O ₂ /kg-solution by Weiss (1970) using the first virial coefficient of oxygen (Atkins 1996, p. 41, 1029) and the atmospheric oxygen fugacity (Williams 2004)
K_W	Millero (1995, p.670) (original reference , but slightly different formula for seawater pH), DOE (1994, chapter 5, p. 18) (NOT the original reference! DOE (1994) cites in an update from 1997 Millero (1995) ! However the version of the formula used here is the one converted to total pH scale given in DOE (1994)), and Zeebe and Wolf-Gladrow (2001, p. 258) . Constant type (stoichiometric), pH scale (total, converted to free here) , and concentration unit (mol/kg-solution squared): DOE (1994, chapter 5, p. 12,18) , pH scale also in Zeebe and Wolf-Gladrow (2001, p. 258) .
K_HSO4	DOE (1994, chapter 5 page 13) , Zeebe and Wolf-Gladrow (2001, p. 260) , Dickson (1990b) (original reference). Constant type (stoichiometric), pH scale (free) , and concentration unit (mol/kg-H ₂ O converted to mol/kg-solution here): DOE (1994, chapter 5, p. 13) . Note that it is also possible to use the constant according to Khoo <i>et al.</i> (1977) , as cited in, e.g., Roy <i>et al.</i> (1993b) , Millero (1995) , and Lewis and Wallace (1998) . In Lewis and Wallace (1998) it is stated that the constant resulting from this equation is in mol/kg-H ₂ O and on the free pH scale.
K_HF	Dickson and Riley (1979b, p. 91) (original reference), DOE (1994, c. 5, p. 15) , Roy <i>et al.</i> (1993b, p. 257) , Dickson and Millero (1987, p. 1783) , Millero (1995, p. 664) , Zeebe and Wolf-Gladrow (2001, p. 260) (converted to molinty and total scale). Constant type (stoichiometric), pH scale (free) , and concentration unit (mol/kg-H ₂ O converted to mol/kg-solution here): DOE (1994, chapter 5, p. 15, 16) . In AquaEnv , it is also possible to use the constant according to Perez and Fraga (1987a) .

K_CO2, K_HCO3	Roy <i>et al.</i> (1993b, p. 254) (original reference), DOE (1994, chapter 5, p.14) (in a version converted to mol/kg-H ₂ O), Millero (1995, p. 664), Zeebe and Wolf-Gladrow (2001, p. 255). Constant type (stoichiometric) and concentration unit (mol/kg-H ₂ O converted to mol/kg-solution here): DOE (1994, chapter 5, p. 14, 15), pH scale (total, converted to free here): In DOE (1994, chapter 5, p. 12) the total scale is stated for the formula for high salinities and thus can be inferred for the formula for low salinities. The scale is also indirectly stated for both formulations in the original reference Roy <i>et al.</i> (1993b). Note that in Roy <i>et al.</i> (1993b) a function for fresh water (based on Millero (1979) which in turn is on a temperature relationship from Harned and Davis (1943) and Harned and Scholes (1941) respectively) and a function for seawater is derived. In Millero (1995) it is stated that for S<5 the fresh water formula (based on Millero (1979)) should be used and for S>=5 the seawater formula derived in Roy <i>et al.</i> (1993b). However, both formulations do not always intersect at S=5. The true intersection with respect to salinity S is a function of temperature t. Here, we first calculate this intersection by numerical root finding and then decide which formulation to use. This practise results in a continuous function with respect to S. (Note that there is a typesetting error in Roy <i>et al.</i> (1993b): one of the numerical values for the function for K _{CO2} [*] is given as 310.48919, but correct is 2310.48919. However, in Millero (1995) this value is stated correctly.) In AquaEnv, it is also possible to use the constants according to Lueker <i>et al.</i> (2000) and Millero <i>et al.</i> (2006).
K_BOH3	Dickson (1990a, p. 763) (original, but mol/kg-H ₂ O version), DOE (1994, ch. 5, p. 14), Zeebe and Wolf-Gladrow (2001, p. 262), Millero (1995, p.669) (mol/kg-H ₂ O version) , agrees with data in Roy, Roy, Lawson, Vogel, Moore, Davis, and Millero (1993a). Constant type (stoichiometric) and concentration unit (mol/kg-solution): DOE (1994, chapter 5, p. 14), pH scale (total): DOE (1994, chapter 5, p. 12) and Zeebe and Wolf-Gladrow (2001, p.263).
K_NH4	Millero <i>et al.</i> (1995) (original reference), Millero (1995, p.671). Constant type (stoichiometric) and concentration unit (mol/kg-solution): Millero (1995, p.671), pH scale (seawater, converted to free here): Lewis and Wallace (1998) (in corrections of Millero (1995)).
K_H2S	Millero <i>et al.</i> (1988) (original reference), Millero (1995, p.671). Constant type (stoichiometric) and concentration unit (mol/kg-solution): Millero (1995, p.671), pH scale (seawater, converted to free here): Lewis and Wallace (1998) (in corrections of Millero (1995)).
K_H3PO4, K_H2PO4, K_HPO4	Millero (1995, p.670) (original reference, but formula for seawater scale pH), DOE (1994, ch. 5, p 16,17), agrees with data in Dickson and Riley (1979a). Constant type (stoichiometric), concentration unit (mol/kg-solution), and pH scale (total, converted to free here): DOE (1994, chapter 5, p. 12, 16, 17).
K_SiOH4	Millero <i>et al.</i> (1988) (original reference), DOE (1994, chapter 5, p 17), Millero (1995, p.671) (formula for seawater scale pH) Constant type (stoichiometric), concentration unit (mol/kg-H ₂ O converted to mol/kg-solution here by omitting the conversion summand ln(1-0.001005 S)), and pH scale (total, converted to free here): DOE (1994, chapter 5, p. 12, 17).
K_SiOOH3	Wischmeyer <i>et al.</i> (2003) (original reference), corrected due to personal communication with Dieter Wolf-Gladrow (one of the authors). The corrected version can be obtained from either Dieter Wolf-Gladrow or Andreas F Hofmann (a.hofmann@nioo.knaw.nl). Constant type (stoichiometric), concentration unit (mol/kg-solution), and pH scale (total, converted to free here): Wischmeyer <i>et al.</i> (2003).
K_HNO2	Constant value, not a function of temperature and salinity! Obtained as a hybrid pk value (featuring the activity of the proton but the concentration of other species (see Zeebe and Wolf-Gladrow (2001) for a treatment of different types of equilibrium constants) in molar concentration (mol/l) on the NBS pH scale (Durst 1975) from Riordan <i>et al.</i> (2005). Used as an approximation for the stoichiometric K _{HNO2} [*] in mol/kg-solution on the free proton pH scale here.
K_H2SO4	Constant value, not a function of temperature and salinity! Obtained as a standard pK value from Atkins (1996, p. 1045). Used as an approximation for the stoichiometric K _{H2SO4} [*] in mol/kg-solution on the free proton pH scale here.
K_HS	Constant value, not a function of temperature and salinity! Obtained as a standard pK value from Atkins (1996, p. 1045). Used as an approximation for the stoichiometric K _{HS} [*] in mol/kg-solution on the free proton pH scale here.
Ksp_calcite, Ksp_aragonite	Mucci (1983) (original reference), Boudreau (1996). Note that in there are errors in Boudreau (1996): b ₀ for calcite is not 0.7712 but 0.77712 and b ₁ for aragonite is not 0.001727 but 0.0017276.

pH	As given in Dickson (1984) , p. 2303 (use of "m") and Dickson and Riley (1979a) , p. 91f all concentrations appearing in the definition of the total and the seawater pH scale are molal (mol/kg-H ₂ O) concentrations. But in Roy et al. (1993b) , p. 257 and in DOE (1994) , chapter 4, SOP 6, p. 1 it is stated, that concentrations for the seawater and total pH scale are in mol/kg-solution. To be consistent with DOE (1994) molal concentrations (mol/kg-solution) are chosen for calculating the pH.
revelle	Zeebe and Wolf-Gladrow (2001, p.73)
dTAdH, dTAdSumCO2, dTAdSumBOH3, dTAdSumH2SO4, dTAdSumHF, dTAdSumH3PO4, dTAdSum- SumSiOH4, dTAdSumH2S, dTAdSumNH4, dTAdSumHNO3, dTAdSumHNO2	Hofmann et al. (2008a)
c1, c2, c3, b1, b2, so1, so2, so3, f1, f2, p1, p2, p3, p4 si1, si2, si3, s1, s2, s3, n1, n2, na1, na2, ni1, ni2	Skoog and West (1982) , Stumm and Morgan (1996) , Hofmann et al. (2008b)
dTAdKdKdS, dTAdKdKdT, dTAdKdKdp, dTAdKdKd- SumH2SO4, dTAdKdKdSumHF	Hofmann et al. (2009)

The values for K_W, K_{HSO4}, K_{HF}, K_{CO2}, K_{HCO3}, K_{BOH3}, K_{NH4}, K_{H2S}, K_{H3PO4}, K_{H2PO4}, K_{HP04}, K_{SiOH4}, K_{SiOOH3}, K_{sp_calcite}, K_{sp_aragonite} obtained as functions of salinity S and temperature t from the above references are pressure corrected using the gauge pressure p according to [Millero \(1995\)](#) with corrections by [Lewis and Wallace \(1998\)](#).

In general it is to be said that all corrections from [Lewis and Wallace \(1998\)](#) have been applied.

References

- Anderson LG, Turner DR, Wedborg M, Dyrssen D (1999). "Determination of total alkalinity and total dissolved inorganic carbon." In K Grasshoff, K Gremling, M Ehrhardt (eds.), "Methods of Seawater Analysis," Wiley-VCH.
- Atkins PW (1996). *Physikalische Chemie*. VCH Weinheim, 2nd edition.
- Boudreau BP (1996). "A method-of-lines code for carbon and nutrient diagenesis in aquatic sediments." *Computers & Geosciences*, **22**(5), 479–496. URL [<GotoISI>://A1996UQ20500003](#).

- Dickson AG (1981). "An Exact Definition of Total Alkalinity and a Procedure for the Estimation of Alkalinity and Total Inorganic Carbon from Titration Data." *Deep-Sea Research Part a-Oceanographic Research Papers*, **28**(6), 609–623. URL [GotoISI://A1981LY51900007](https://doi.org/10.1016/0198-0178(81)90007-1).
- Dickson AG (1984). "Ph Scales and Proton-Transfer Reactions in Saline Media Such as Sea-Water." *Geochimica Et Cosmochimica Acta*, **48**(11), 2299–2308. URL [GotoISI://A1984TT08000013](https://doi.org/10.1016/0016-7136(84)90013-1).
- Dickson AG (1990a). "Standard Potential of the Reaction - $\text{AgCl(S)} + 1/2\text{H}_2(\text{G}) = \text{Ag(S)} + \text{HCl(Aq)}$ and the Standard Acidity Constant of the Ion HSO_4^- in Synthetic Sea-Water from 273.15-K to 318.15-K." *Journal of Chemical Thermodynamics*, **22**(2), 113–127. URL [GotoISI://A1990CY50200001](https://doi.org/10.1016/0360-5310(90)90001-1).
- Dickson AG (1990b). "Thermodynamics of the Dissociation of Boric-Acid in Synthetic Seawater from 273.15-K to 318.15-K." *Deep-Sea Research Part a-Oceanographic Research Papers*, **37**(5), 755–766. URL [GotoISI://A1990DK56300004](https://doi.org/10.1016/0198-0178(90)90004-1).
- Dickson AG, Millero FJ (1987). "A Comparison of the Equilibrium-Constants for the Dissociation of Carbonic-Acid in Seawater Media." *Deep-Sea Research Part a-Oceanographic Research Papers*, **34**(10), 1733–1743. URL [GotoISI://A1987L675400008](https://doi.org/10.1016/0198-0178(87)90008-1).
- Dickson AG, Riley JP (1979a). "Estimation of Acid Dissociation-Constants in Seawater Media from Potentiometric Titrations with Strong Base .1. Ionic Product of Water - K_w ." *Marine Chemistry*, **7**(2), 89–99. URL [GotoISI://A1979GJ26800001](https://doi.org/10.1016/0304-3866(79)90001-1).
- Dickson AG, Riley JP (1979b). "Estimation of Acid Dissociation-Constants in Seawater Media from Potentiometric Titrations with Strong Base .2. Dissociation of Phosphoric-Acid." *Marine Chemistry*, **7**(2), 101–109. URL [GotoISI://A1979GJ26800002](https://doi.org/10.1016/0304-3866(79)90002-1).
- Dickson AG, Sabine C, Christian JR (2007). "Guide to best practices for ocean CO₂ measurements." *PICES special publications*, (3), 1–191.
- DOE (1994). *Handbook of Methods for the Analysis of the Various Parameters of the Carbon Dioxide System in Sea Water*. ORNL/CDIAC-74.
- Durst A (1975). *Standard Reference Materials: Standardization of pH Measurements*, volume 260-53 of *NBS Spec. Publ.* National Bur. Standards, Washington, D.C.
- Emerson S, Hedges JI (2008). "Carbonate Chemistry." In "Chemical Oceanography and the Marine Carbon Cycle," pp. 101–133. Cambridge University Press, Cambridge.
- Feistel R (2008). "A Gibbs function for seawater thermodynamics for -6 to 80 degrees C and salinity up to 120 g kg⁻¹." *Deep-Sea Research Part I-Oceanographic Research Papers*, **55**(12), 1639–1671. URL [GotoISI://000261725600003](https://doi.org/10.1016/j.dsr.2008.09.003).
- Fofonoff NP, Millard RCJ (1983). "Algorithms for computation of fundamental properties of seawater." *Unesco technical papers in marine science*, **44**, 55 pp.
- Follows MJ, Ito T, Dutkiewicz S (2006). "On the solution of the carbonate chemistry system in ocean biogeochemistry models." *Ocean Modelling*, **12**(3-4), 290–301. URL [GotoISI://000235695500003](https://doi.org/10.1016/j.oceanmod.2006.05.003).

- Gran G (1952). "Determination of the Equivalence Point in Potentiometric Titrations .2." *Analyst*, **77**(920), 661–671. URL [GotoISI://A1952UG08300014](https://doi.org/10.1039/JR9520000661).
- Hansson I, Jagner D (1973). "Evaluation of Accuracy of Gran Plots by Means of Computer Calculations - Application to Potentiometric Titration of Total Alkalinity and Carbonate Content in Sea-Water." *Analytica Chimica Acta*, **65**(2), 363–373. URL [GotoISI://A1973Q206600014](https://doi.org/10.1016/0003-2700(73)90014-1).
- Haraldsson C, Anderson LG, Hasselov M, Hulth S, Olsson K (1997). "Rapid, high-precision potentiometric titration of alkalinity in ocean and sediment pore waters." *Deep-Sea Research Part I-Oceanographic Research Papers*, **44**(12), 2031–2044. URL [GotoISI://000072760700007](https://doi.org/10.1016/S0967-6266(97)00007-1).
- Harned HS, Davis R (1943). "The Ionization Constant of Carbonic Acid in Water and the Solubility of Carbon Dioxide in Water and Aqueous Salt Solutions from 0 to 50 deg C." *Journal of the American Chemical Society*, **65**(10), 2030–2037. URL [http://pubs.acs.org/doi/abs/10.1021/ja01250a059](https://pubs.acs.org/doi/abs/10.1021/ja01250a059).
- Harned HS, Scholes SR (1941). "The Ionization Constant of HCO₃⁻ from 0 to 50 deg C." *Journal of the American Chemical Society*, **63**(6), 1706–1709. URL [http://pubs.acs.org/doi/abs/10.1021/ja01851a058](https://pubs.acs.org/doi/abs/10.1021/ja01851a058).
- Hofmann A, Meysman F, Soetaert K, Middelburg J (2009). "Factors governing the pH in a heterotrophic, turbid, tidal estuary." *Biogeosciences Discussion*, **6**, 197–240.
- Hofmann AF, Meysman FJR, Soetaert K, Middelburg JJ (2008a). "A step-by-step procedure for pH model construction in aquatic systems." *Biogeosciences J1 - BG*, **5**(1), 227–251. URL <http://www.biogeosciences.net/5/227/2008/L1-http://www.biogeosciences.net/5/227/2008/bg-5-227-2008.pdf>.
- Hofmann AF, Middelburg J, Soetaert K, Wolf-Gladrow DA, Meysman F (2008b). "pH modelling in aquatic environments: from an alkalinity-based to a proton-based view." *in preparation*.
- Khoo KH, Ramette RW, Culberson CH, Bates RG (1977). "Determination of hydrogen ion concentrations in seawater from 5 to 40 °C: standard potentials at salinities from 20 to 45 ‰." *Analytical Chemistry*, **49**, 29–34.
- Lewis EL, Wallace DWR (1998). "Program Developed for CO₂ System Calculations." ORNL/CDIAC-105. Carbon Dioxide Information Analysis Center, Oak Ridge National Laboratory, U.S. Department of Energy, Oak Ridge, Tennessee.
- Lueker TJ, Dickson AG, Keeling CD (2000). "Ocean pCO₂ calculated from dissolved inorganic carbon, alkalinity, and equations for K₁ and K₂: validation based on laboratory measurements of CO₂ in gas and seawater at equilibrium." *Marine Chemistry*, **70**(1-3), 105–119. URL [GotoISI://000087534100009](https://doi.org/10.1016/S0304-4203(99)00009-1).
- Millero FJ (1979). "Thermodynamics of the Carbonate System in Seawater." *Geochimica Et Cosmochimica Acta*, **43**(10), 1651–1661. URL [GotoISI://A1979HP85900006](https://doi.org/10.1016/0016-7037(79)90006-1).
- Millero FJ (1982). "The Thermodynamics of Seawater, Part 1. The PVT Properties." *Ocean Science and Engineering*, **7**(4), 403–460.

- Millero FJ (1995). "Thermodynamics of the Carbon-Dioxide System in the Oceans." *Geochimica Et Cosmochimica Acta*, **59**(4), 661–677. URL [GotoISI://A1995QH99100003](https://doi.org/10.1016/0016-7037(95)00003-1).
- Millero FJ, Graham TB, Huang F, Bustos-Serrano H, Pierrot D (2006). "Dissociation constants of carbonic acid in seawater as a function of salinity and temperature." *Marine Chemistry*, **100**(1-2), 80–94.
- Millero FJ, Plese T, Fernandez M (1988). "The Dissociation of Hydrogen-Sulfide in Seawater." *Limnology and Oceanography*, **33**(2), 269–274. URL [GotoISI://A1988N101700008](https://doi.org/10.4319/limnol.1988.33.2.0269).
- Millero FJ, Poisson A (1981). "International One-Atmosphere Equation of State of Seawater." *Deep-Sea Research Part a-Oceanographic Research Papers*, **28**(6), 625–629. URL [GotoISI://A1981LY51900008](https://doi.org/10.1016/0198-0149(81)90008-1).
- Millero FJ, Yao WS, Aicher J (1995). "The Speciation of Fe(II) and Fe(III) in Natural Waters." *Marine Chemistry*, **50**(1-4), 21–39. URL [GotoISI://A1995RV93400004](https://doi.org/10.1016/0304-4203(95)00004-1).
- Mucci A (1983). "The Solubility of Calcite and Aragonite in Seawater at Various Salinities, Temperatures, and One Atmosphere Total Pressure." *American Journal of Science*, **283**(7), 780–799. URL [GotoISI://A1983RF42100006](https://doi.org/10.2475/ajsc.283.7.780).
- Perez FF, Fraga F (1987a). "Association Constant of Fluoride and Hydrogen-Ions in Seawater." *Marine Chemistry*, **21**(2), 161–168. URL [GotoISI://A1987J311800005](https://doi.org/10.1016/0304-4203(87)90005-1).
- Perez FF, Fraga F (1987b). "The Ph Measurements in Seawater on the Nbs Scale." *Marine Chemistry*, **21**(4), 315–327. URL [GotoISI://A1987K470200002](https://doi.org/10.1016/0304-4203(87)90002-1).
- Riordan E, Minogue N, Healy D, O'Driscoll P, Sodeau JR (2005). "Spectroscopic and optimization modeling study of nitrous acid in aqueous solution." *Journal of Physical Chemistry A*, **109**(5), 779–786. URL [GotoISI://000226779800009](https://doi.org/10.1021/jp0426779).
- Roy RN, Roy LN, Lawson M, Vogel KM, Moore CP, Davis W, Millero FJ (1993a). "Thermodynamics of the Dissociation of Boric-Acid in Seawater at S=35 from 0-Degrees-C to 55-Degrees-C." *Marine Chemistry*, **44**(2-4), 243–248. URL [GotoISI://A1993MT06600012](https://doi.org/10.1016/0304-4203(93)90012-1).
- Roy RN, Roy LN, Vogel KM, PorterMoore C, Pearson T, Good CE, Millero FJ, Campbell DM (1993b). "The dissociation constants of carbonic acid in seawater at salinities 5 to 45 and temperatures 0 to 45 degrees C (vol 44, pg 249, 1996)." *Marine Chemistry*, **52**(2), 183–183. URL [GotoISI://A1996UH75200007](https://doi.org/10.1016/0304-4203(96)00007-1).
- Skoog DA, West DM (1982). *Fundamentals of Analytical Chemistry*. Holt-Saunders International Editions. Holt-Saunders.
- Stumm W, Morgan JJ (1996). *Aquatic Chemistry: Chemical Equilibria and Rates in natural Waters*. Wiley Interscience, New York.
- Sundquist ET, Plummer LN, Wigley TML (1979). "Carbon-Dioxide in the Ocean Surface - Homogeneous Buffer Factor." *Science*, **204**(4398), 1203–1205. URL [GotoISI://A1979GY02300024](https://doi.org/10.1126/science.1203120).
- Weiss RF (1970). "Solubility of Nitrogen, Oxygen and Argon in Water and Seawater." *Deep-Sea Research*, **17**(4), 721–735. URL [GotoISI://A1970H126100004](https://doi.org/10.1016/0198-0149(70)90004-1).

- Weiss RF (1974). “Carbon dioxide in water and seawater: the solubility of a non-ideal gas.” *Marine Chemistry*, **2**, 203–215.
- Williams DR (2004). “NASA Earth Fact Sheet.” URL <http://nssdc.gsfc.nasa.gov/planetary/factsheet/earthfact.html>.
- Wischmeyer AG, Del Amo Y, Brzezinski M, Wolf-Gladrow DA (2003). “Theoretical constraints on the uptake of silicic acid species by marine diatoms.” *Marine Chemistry*, **82**(1-2), 13–29. URL [GotoISI://000183605000002](http://www.sciencedirect.com/science/article/pii/S030438650000002).
- Zeebe RE, Wolf-Gladrow D (2001). *CO₂ in Seawater: Equilibrium, Kinetics, Isotopes*. Number 65 in Elsevier Oceanography Series. Elsevier, first edition.

Affiliation:

Andreas F. Hofmann
Centre for Estuarine and Marine Ecology (CEME)
Netherlands Institute of Ecology (NIOO)
4401 NT Yerseke, Netherlands E-mail: a.hofmann@nioo.knaw.nl
URL: <http://www.nioo.knaw.nl/users/ahofmann>